

**Biomass/carbon estimation and mapping
in the subtropical forest of Chitwan, Nepal:
A comparison between VHR GeoEye
satellite images and airborne LiDAR data**

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February, 2012

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ABSTRACT

It is generally agreed that preservation of forest areas can contribute strongly to the mitigation of global climate change. For this purpose, international institutions such as the United Nations Framework on Climate Change (UNFCCC) have created a collaborative program in reduction emissions of carbon dioxide (REDD) to update inventories emissions from greenhouse gasses. However, studies have demonstrated that there are still uncertainties for an accurate estimation of carbon stock from forests, especially using optical remote sensing. This study aims therefore to determine which of the two sources from airborne LiDAR data or VHR GeoEye satellite images can provide more accurate information for biomass/carbon estimation in the Subtropical forest of Chitwan, Nepal.

A very high resolution GeoEye satellite image provides information only in two dimensions while LiDAR data provides information in three dimensions. In the approach of this study, LiDAR data required more analysis because original information from the sensor is acquired in a cloud of points. Then, a Digital Surface Model (DSM) and a Digital Terrain Model (DTM) were derived from the cloud of points. Canopy Height Model (CHM), which is the height of trees, was calculated as the difference between DSM and DTM. The height of the trees derived from LiDAR data were compared to the height of trees measured in the field.

LiDAR CHM and GeoEye images were segmented using the technique of Object Oriented Analysis (OOA) to delineate individual tree crowns and the results were compared with the manual delineation derived from the field. Then, the segments derived from both images were used to develop models to estimate the Diameter at breast height (DBH). With the most accurate DBH estimated, Above Ground Biomass (AGB) was calculated using an allometric equation that considers DBH, Height (H) and Wood specific gravity (ρ).

Results show that there is no significant difference between the height of trees derived from LiDAR data and the height of trees measured in the field. However, the root mean square error (RMSE) shows a relative value of 27%. The result from segmentation implies that objects from GeoEye are oversegmented compared to those derived from LiDAR data with a difference of 14%. Segments derived from GeoEye and LiDAR images were used to develop models and shows that power model and LiDAR data is more accurate to predict DBH from CPA. Finally, the estimation of carbon stock resulted in a mean value of 1894.08 kg C/tree, which is equivalent to 181.34 Mg C ha⁻¹. Thus, biomass/carbon estimation and mapping in the subtropical forest is practicable utilizing LiDAR data.

Keywords: *Crown projection area, Canopy height model, Object Oriented Analysis, Allometric equation, Regression, Carbon stock*

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LIST OF ACRONYMS

| | |
|-----------------|--|
| IPCC | Intergovernmental Panel on Climate Change |
| UNFCCC | United Nations Framework Convention on Climate Change |
| FAO | Food Agricultural Organization |
| REDD | Reducing Emission from Deforestation and Degradation |
| GHG's | Greenhouse gases |
| CO ₂ | Carbon dioxide |
| ICIMOD | International Centre for Integrated Mountain Development |
| ANSAB | Asia Network for Sustainable Agriculture and Bio-resources |
| CFUG | Community forest user group |
| GPS | Geographic Position System |
| VHR | Very high resolution |
| DTM | Digital Terrain Model |
| DSM | Digital Surface Model |
| MSS | Multispectral data |
| AGB | Aboveground biomass |
| CPA | Crown projection area |
| DBH | Diameter at breast height |
| OOA | Object oriented analysis |
| RMSE | Root Mean Square Error |

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1. INTRODUCTION

1.1. Background

Climate change is a product of Green House Gas (GHG) emissions associated with the provision of energy services causing the current global warming (IPCC, 2011). Gases that contribute to the greenhouse effect are: Water vapour, Carbon dioxide (CO₂), Methane (CH₄), Nitrous oxide (N₂O) and Chlorofluorocarbons (CFCs) (IPCC, 1990). It is considered that annual emissions of CO₂ grew by about 80% between 1970 and 2004, while the global atmospheric concentration increased from 280 ppm to 379 ppm³ by 2005, and it has been indicated that the annual growth rate of CO₂ has increased by 1.9 ppm y⁻¹ between 1995 and 2005 (IPCC, 2007b; McKibben, 2007). The emission of gasses is a product of natural processes such as volcanic eruptions or mainly by human activities, including: deforestation, land use changes, burning fossil fuels, and agriculture like soil cultivation practices.

Deforestation is one of those processes that deserve attention, given that forest can be considered as *'double'* significant in combating global warming. This is because: (1) deforestation adds CO₂ to the atmosphere when the carbon contained in the forest is burnt or decomposed, and (2) increases in forest biomass extracts CO₂ from the atmosphere and stores it. Carbon is stored by trees in their roots, trunks, branches and leaves. The removal of 3.67 tonnes of carbon dioxide from the atmosphere results in one tonne of carbon sequestered in trees (Hunt, 2009). In the Global Forest Resource Assessment of FAO (2010), it is estimated that all carbon stored in above ground biomass, litter and soils of the entire worlds' forests is around 652 billion tonnes with an average carbon content of 161.8 tonnes per hectare. The UN's Intergovernmental Panel on Climate Change (IPCC, 2007a) estimated in 2007 that deforestation, forest degradation and other changes in forests, contributes in 17.4 per cent of global greenhouse emissions. Then, Vegetation biomass is an important ecological variable for understanding the changes of climate and carbon sequestration in a forest (FAO, 2009). It is generally agreed that preservation of forest areas can contribute strongly to the mitigation of global climate change. For this purpose, the United Nations Framework on Climate Change (UNFCCC) have created a collaborative program on 2008 called Reducing Emissions from Deforestation and Forest Degradation (REDD) to assist developing countries to build capacity in reduction emissions to future involvement on REDD+ mechanism. Accordingly, REDD+ refers to reduce these emissions from deforestation and forest degradation and the role of conservation, sustainable management of forests and enhancement of forest carbon stocks (UN-REDD, 2008a). For this reason REDD+ facilitates the provision of financial incentive to developing countries by reducing emissions from forested lands and investing in low-

carbon paths for sustainable development (UN-REDD, 2008b). In order to appropriate monitoring systems of forest lands it is necessary to combine different data sources such as: field measurements and observations, varieties of satellite and airborne remote sensing data with cost-efficiency to achieve carbon reports with an accepted accuracy and precision. Remote sensing data is a crucial source of information for monitoring the state and trends of land-use, land cover and carbon estimation in a particular area (Holmgren, 2008).

1.2. Methods for above ground biomass estimations

Many methods have been used in forest biomass estimation. The most accurate method for estimation of forest biomass is to harvest the trees, oven-dry them and to weigh the dry matter. However, this direct method is normally prohibitively expensive, destructive and time-consuming (Hunt, 2009). In recent years, many studies using satellite images with high resolution (Kosaka & Kuwata, 2006) and radar have been investigated as alternative methods for biomass estimation (Patenaude et al., 2005). These methods have been explored in terms of biomass and its changes to increase our understanding of the role of forests in the carbon cycle for greenhouse gas inventories and terrestrial carbon accounting (Muukkonen & Heiskanen, 2007). For example, as studied by Dubayah et al. (2000), information such as canopy cover, tree size (height and crown diameter), biomass, crown volume among others are routinely needed for sustainable resource management because it is useful information that can help to maintain and enhance the economic, social and environmental value of all types of forests, for the benefit of present and future generations. Despite the fact that optical images with very high spatial resolution, such as GeoEye, can provide useful information of spectral reflectance in objects, there is still information like vertical parameters (height of trees) that cannot be collected with this sensor and can be useful to improve biomass estimations. In addition, weather conditions can affect the image quality, especially in those areas that are frequently covered by clouds. This is a situation that occurs frequently in many tropical and subtropical areas.

Recently, Light Detection and Ranging (LiDAR) is becoming more a promising technique for future forest monitoring because of its ability to assess the 3D forest structure (Patenaude et al., 2005; van Leeuwen & Nieuwenhuis, 2010) and to provide a good data on vertical profiles of vegetation canopies (Balzter et al., 2007). LiDAR is an active remote sensing system which operates from aircraft by sending laser pulses towards the ground to record the elapsed time between beam launch and return registration. Records are a cloud of points product of pulses reflected from tree canopy, trunks, branches, leaves, low vegetation and even reaching to the ground to create a profile in three dimensions (Gautam & Kandel, 2010). In this way, the vertical parameters of individual trees are calculated, giving the most feasible advantage of LiDAR to assess biomass and carbon estimation (Kim & et al, 2010). Information of vertical profiles in individual trees can be used to validate models

from other sensors with accurate values (Balzter et al., 2007) or to estimate biomass in dense tropical forests (Drake et al., 2002). Therefore, they have received increasing attention as new technology to derive forest structural attributes. This can be provided and differentiated by direct retrieval (as canopy height and crown volume), modelling (as vertical foliar diversity and layers in ecological applications) or by fusion with other sensors (such as vegetation type). Considering these advantages, traditional multispectral classifications in forest may be more accurately estimated when vertical component provided by LiDAR is added (Dubayah et al., 2000).

The greatest advantage of LiDAR is its capacity to assess forest degradation when it is combined with advanced statistical models calibrated with sample plots. Whereas other optical remote sensing only provide the highest layer of the canopy suffering from saturation problems, while LiDAR beams always penetrates into the deep forest and the resulting data are without being influenced by clouds and shadows, thus providing more accurate results than any other remote sensing techniques (ARBONAUT, 2010; Næsset, 2009).

1.2.1. Biomass estimation using Height and Diameter at Breast Height (DBH)

Different parameters can be used for above ground biomass estimation, like basal area, stand structure and tree height. However, DBH is the widely available and is most commonly used for this estimation (Crow & Schlaegel, 1988). Since DBH of trees cannot be derived directly from RADAR, LiDAR or optical remote sensing data, regression models have been developed based on crown projection area (CPA) to assess AGB (Chave et al., 2005) (See also section 2.2).

Also, canopy height provides an opportunity to model ABG and canopy volume. It has been shown in an accurate way that is possible to determine the height of trees using remote sensing like multiparametric interferometric radar (like Pol-InSAR) and more recently using LiDAR data. As a comparison of these two methods, Pol-InSAR does not retrieve heights directly, but indirectly by applying a model-based inversion (Mette et al., 2004). LiDAR data can directly retrieve tree height data (Lim et al., 2003), offering a new way to describe a forest structure in 3D (Maier et al., 2008).

1.3. The study area

In 2008, the Nepal government submitted a note to the United Nations Framework Convention on Climate Change (UNFCCC) Committee to be considered under REDD project and to have economical financing on forest carbon stock. In the same year Nepal was selected as a winner for Forest Carbon Partnership Facility (FCPF) fund, to develop the Readiness Plan -R-PLAN- (Dahal & Banskota, 2009) to undertake activities on reducing emissions for deforestation and forest degradation in the REDD context. The REDD context refers to establishing scenarios for emission reductions from deforestation

and forest degradation and establish scenarios in order to reduce emissions. To address international issues to national context is becoming a challenge in Nepal, mainly by the weak governance and to be in a process of state restructuring. Actually, this country has produced sufficient data in terms of satellite images and field data which is required for biomass estimation to fulfil the REDD context. This was also taken into consideration in a project by the International Centre for Integrated Mountain Development (ICIMOD) (Dahal & Banskota, 2009), that aims to assist mountain people to understand changes induced by climate and to adapt them to the changes and make the most of new opportunities. Afterwards, due to the benefits mentioned above, most of the forests became managed by communities and are considered as models of management to reduce its degradation in the Mid-hills of Nepal. This contributes to the enhancement of carbon stock (Ministry of Forest and Soil Conservation, 2010).

As continuation of the project in Nepal, many study projects have been developed for biomass/carbon estimation in different mountainous areas. Recently, Baral (2011) has mapped carbon stock using high resolution satellite images in the same forest as in this study obtaining a relative error of 39%. She concluded that subtropical regions like Chitwan are fast growing and diverse, emphasizing that intermingling between canopies of individual trees can affect the carbon stock estimation.

LiDAR data has been also used to estimate above ground biomass in mixed species and result obtained by Garcia, et al., (2010) shows a root mean square error (RMSE) in a range of 9.7 Mg ha⁻¹ and 18.48 Mg ha⁻¹, emphasizing that LiDAR data can estimate carbon content in mixed species with high accuracy.

In the case of Nepal, it is a significant challenge to use conventional remote sensing for biomass estimation due to the complexity of the terrain, increasing the possibility of LiDAR to assess the forest resources. Processes of factors that are influencing the biomass/carbon stock and methods to its estimation are presented in Figure 1. Environmental factors and policies can affect or benefit the forests, however, analyses in forests need to be assessed to estimate the carbon stock and to contribute to the mitigation of carbon dioxide emissions and climate change. Methods using remote sensing techniques are common and feasible, especially in remote areas.

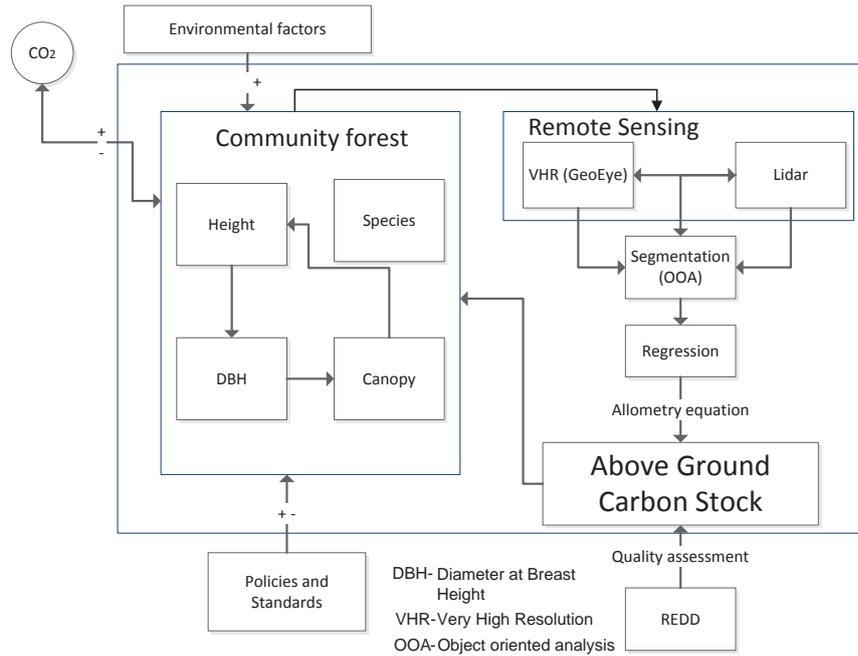


Figure 1 Theoretical framework

This research was carried out in a study area falling in the subtropical forest located in Chitwan District, Nepal (Figure 2). This is a forest managed by the Community (CF) and handed over to a forest user group for its development, conservation and utilization (Ojha et al., 2009).

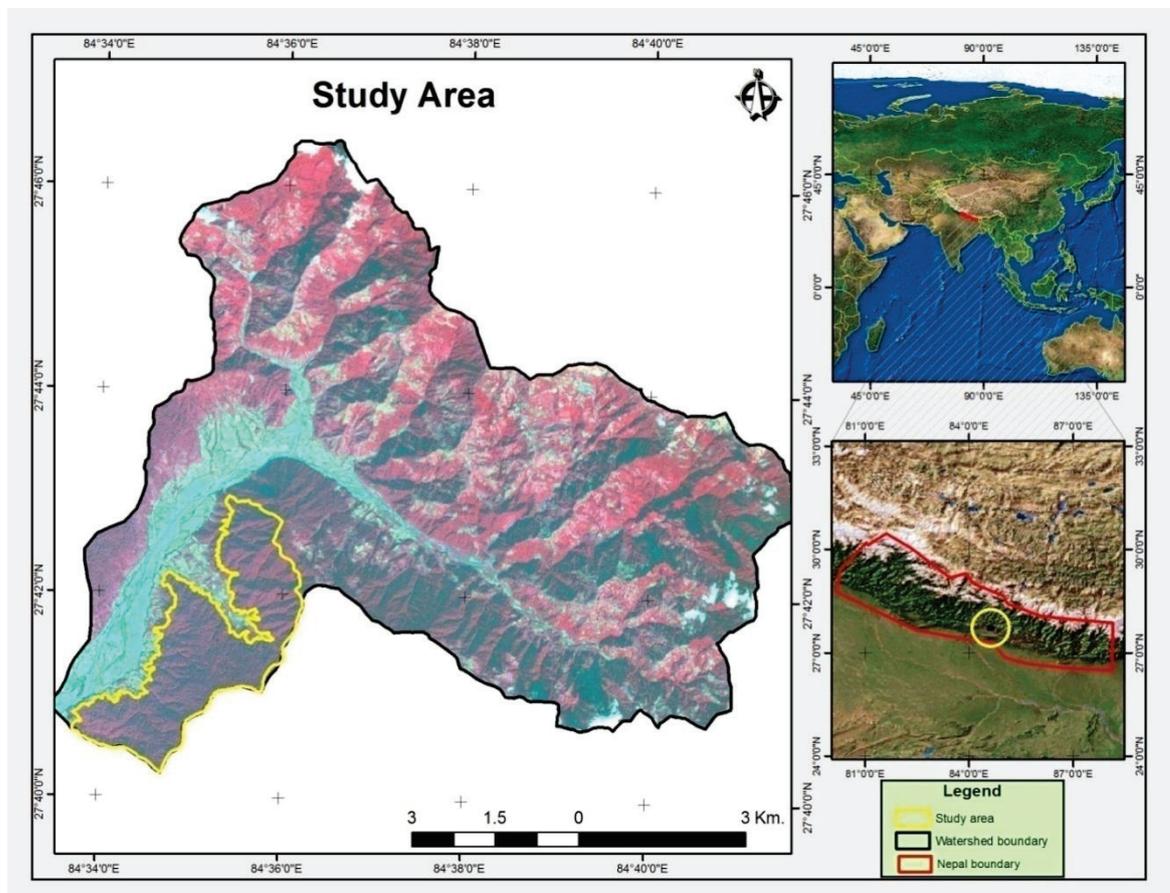


Figure 2 Location of the study area

1.4. Problem Statement

Many efforts have been performed using different methods to estimate biomass with high accuracy to extract reliable structural tree information (Maier et al., 2008). Despite of this, there are still considerable uncertainties in terms of accurate delineation of trees and regarding methods that can standardize and precise these estimations (Nichol & Sarker, 2011). These uncertainties are causing overestimation or underestimation of biomass/carbon, that are being attributed mainly to the complexity of forest ecosystems that add more difficulties to derive forest parameters mainly in those areas located in tropical and subtropical areas (Lu, 2005).

Regarding biomass estimation, Lu (2006) emphasizes the importance to integrate field data with high resolution remote sensing which also agrees with the research of Gautam, et al (2010). Both authors conclude that integration of data provides an accurate, precise and affordable monitoring solution for tropical forests. However, field data is time consuming and labour intensive in remote areas (Brown, 2002; Houghton et al., 2001). Then, VHR images have been widely applied for biomass estimation based on individual trees, demonstrating correct detection of trees for biomass estimations (Nichol & Sarker, 2011). Despite of this, Tsendbazar (2011) found that effect of shadows using VHR optical images can have influence to find a relationship between CPA and carbon stock of trees, because shadow can influence in the geometry of tree crowns which makes it difficult to distinguish crowns using remotely sensed images.

More recently, LiDAR from satellite platforms has become available and has been widely applied in biomass estimations (Patenaude et al., 2005). Its advantage is based in its ability to provide height values of trees that can be added to these estimations. Kim et al. (2010) was using LiDAR data to deal with the difficulties and error on carbon storage in case of high tree density, where crown overlapping occurs in a forest of *Pinus desinflora* located in South Korea. It is mentioned that error can be reduced when height of trees are added to identify gap areas via the height information of tree stands using LiDAR data. These advantages on LiDAR data are results of its potential in providing 3D information to characterize forest for biomass estimation (van Leeuwen & Nieuwenhuis, 2010).

This study focuses on the biomass/carbon estimation using Object Oriented Analysis (OOA) as basis to delineate individual trees. OOA is a technique widely discussed in processing VHR satellite images (Heyman et al., 2003; Yu et al., 2006), used to delineate and extract individual trees inside dense subtropical forests for biomass/carbon estimation. Considering this advantage, individual trees are delineated inside dense subtropical forests using airborne LiDAR data and VHR GeoEye satellite images to determine the most appropriate method in the delineation of individual trees in comparison

with the field data. Then, the most accurate data source (LiDAR or GeoEye) in the delineation of individual trees is used for biomass/carbon estimation.

1.5. Research Objectives

This study aims to determine which of two different sources - airborne LiDAR data and VHR GeoEye satellite images - can provide more accurate information for biomass/carbon estimation in the Subtropical forest of Chitwan, Nepal.

1.5.1. The Specific Objectives

1. To analyze the relationship between tree height measured in the field and tree height values derived from LiDAR data
2. To estimate the segmentation accuracy of OOA on LiDAR data compared to segmentation of VHR GeoEye data
3. To determine the relationship between Crown Projection Area (CPA) and Diameter at Breast Height (DBH)
4. To map the estimated biomass/carbon in the Subtropical forest of Chitwan, Nepal, using the best performing method found from the results of objective 2.

1.6. Research questions and hypothesis

| Objectives | Research Questions | Hypothesis |
|------------|---|--|
| 1 | How strong is the relationship between tree heights measured in the field with height values derived from LiDAR data? | H1. There is a strong significant relationship between tree height values measured in the field and tree height values derived from LiDAR data |
| 2 | Does image segmentation from LiDAR data give an improvement in accuracy compared to GeoEye images? | H1. Image segmentation from LiDAR data has an improvement compared to VHR GeoEye images |
| 3 | How strong is the relationship between CPA and DBH? | H1. There is a strong significant relationship between CPA and DBH |

2. CONCEPTS AND DEFINITIONS

2.1. Biomass and Carbon

Biomass concerns the dry weight of the trees. Generally it includes the above ground and below ground living mass, like trees, shrubs, vines, roots and the dead mass of fine and coarse litter associated with the soil (Lu, 2006). To quantify biomass, the most appropriate and direct way is to harvest all trees of a certain area, dry them and weigh the biomass (Gibbs, 2007). Then, most commonly carbon content in a forest can be estimated by multiplying dry biomass by a fraction of 0.5 (IPCC, 2006), while after careful comparison ICIMOD (2010) has proposed to convert biomass value of standing trees into Carbon stock multiplying by the factor 0.47 for forests in Nepal.

2.2. Remote sensing approach in estimating carbon stock

Traditional techniques, based on field measurements (Brown, 2002; Houghton et al., 2001) are most accurate, but are time consuming and labour intensive, especially in remote areas. Also geographic Information Systems (GIS) was used by Brown & Gaston (1995) in the tropical forest with the use of available spatial data sets to extend limited biomass data to regional and national estimations. Concluding that it is difficult to obtain good quality of data and that also has not been used widely for above ground biomass (AGB) estimations. These disadvantages on GIS and traditional techniques are providing more advantages to remote sensing techniques, increasing the attraction of scientific interest (Foody et al., 2003; Lu et al., 2005; Santos et al., 2003; Zheng et al., 2004).

In this sense, RS has become widely used and a solution to AGB estimation (Lu, 2005; Steininger, 2000; Zheng et al., 2004). However, most of the studies related to –AGB- have been focused on coniferous forests due to the species composition and stand structure which is relatively easy to study, compared to moist tropical forests that become complex due to the wide species composition and complex stand structure (Lu et al., 2005; Zheng et al., 2004).

2.2.1. Above-ground biomass estimation with Optical Sensor Data

Different approaches have been used to estimate AGB based on remote sensed data, like crown diameter using regression analysis to estimate DBH or using canopy reflectance models (Phua & Saito, 2003; Popescu et al., 2003). Also others, such as neural network, K nearest-neighbour (Foody et al., 2003; Nelson et al., 2000; Zheng et al., 2004) have been less used. Thereupon, sensors from very high to low spatial resolution have been used for AGB estimation. This section summarizes the AGB estimation based on very high spatial resolution (VHSR) of optical remote sensing.

Very high spatial resolution data (VHSR)

VHSR also called fine spatial resolution data can be airborne such as aerial photographs or space-borne such as Ikonos, GeoEye, Quickbird images with a spatial resolution lower than 5 meters. These data types have been frequently used for modelling tree parameters (Levesque & King, 2003). For applications related to forest inventory, aerial photograph has become extensively used, considering that when using photo-interpretation technique and stereo pairs it is possible to measure various forest parameters, like tree height, crown diameter and stand area. Nevertheless, since VHSR optical images are providing similar information (except height) and contain a similar spatial and spectral resolution, these have been increasingly used for forest inventory studies (Wulder et al., 2004). Culvenor (2003) evaluates the extraction of individual trees using fine spatial resolution images and concludes that achieving information from forest inventories is still a challenge because of the complex structure of natural forest canopies. Nichol & Sarker (2011) improved the biomass estimation in a study located in China using two high resolution optical sensors (Advanced Visible and Near Infrared Radiometer type 2 (AVNIR-2) and SPOT 5). They developed multiple regression models between information derived from the images and biomass derived from field data and results show an r^2 of 0.65. Tiede et al., (2008) combined airborne laser scanner (ALS) with optical image data to perform automated tree crown delineation. The method was based on Cognition Network Language which is purely a programming language. Results demonstrate correctly detected accuracies of trees (concerning the location of trees) between 86 % for coniferous and 79 % for dead trees, but dropping to 44 % for deciduous trees.

2.3. Above-ground biomass estimation with Radar and LiDAR data

In tropical areas, where the climatic conditions, especially clouds, are affecting most of the year, the use of optical sensors becomes difficult. Thus, radar data become a feasible way for forest studies. Previous work has identified the potential of Radar in estimating AGB (Hussin et al., 1991; Santos et al., 2003). Hussin, et al., (1991) has concluded that when using L-band HV there is a strong positive relationship with pine stand parameters (biomass, height and basal area) and Santos, et al., (2003) found that using polarimetric P-band data can contribute significantly to develop models in tropical areas where it is difficult to obtain information from optical remote sensing. Furthermore, the fusion of Radar with optical remote sensing is a potential technique because it enables the possibility to have vertical attributes of forests together with 2D information from optical data (Treuhaft et al., 2004). In the same way, LiDAR data has shown a promising approach for biophysical parameter estimation and AGB assessment. This is mainly due to the advantage of not being affected by clouds and its ability in providing information in 3D (Drake et al., 2002; Hyde et al., 2005).

2.4. Carbon stock from Crown Projection Area (CPA)

CPA also known as canopy cover is defined as the area of the ground which is reflected by a vertical projection of the tree crowns (Jennings et al., 1999). Field measurements of CPA can be difficult because of the irregularity of the tree crown's outline. To facilitate the measurement, the usual way is to project the perimeter of the crown vertically to the ground and to make a diameter measurement taking an average of two perpendicular direction of the canopy diameter (Husch et al., 2003) (*Figure 3*).

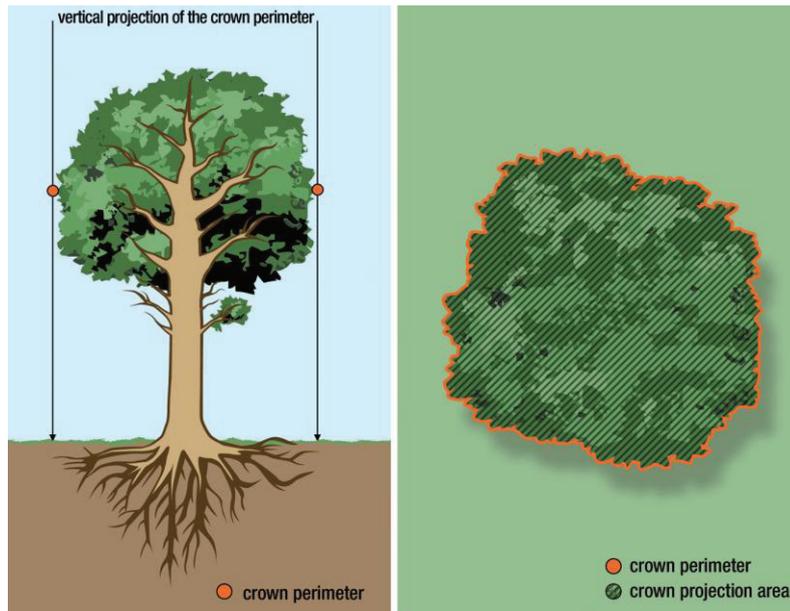


Figure 3 Crown projection area (Gschwantner et al., 2009)

2.5. Allometric Equation

Ketterings, et al. (2001) define allometric equation as a quantitative relationship between measurable tree variables like DBH and height to other difficult to assess variables, like standing volume of the tree or total biomass or carbon stock. For biomass estimations Chave, et al. (2005) emphasize that biomass regression models may include information on diameter at breast height DBH (1.3 m from the ground), total tree height H (in m) and wood specific gravity ρ (in g/cm^3) and developed an allometric equation for biomass estimation with these three parameters and found a bias of 0.5-6.5%. While Brown (2002) shows that for reliable carbon estimations in highly diverse tropical forests, allometric equations can be used that consider only DBH measurements.

2.6. Object Oriented Analysis (OOA)

OOA also called object based image analysis (OBIA) refers to the partition of an image into discrete non overlapping units called image objects considering the homogeneity in terms of spectral or spatial properties (Hay et al., 2005; Zhang et al., 2010).

Various OOA image classification techniques have been used successfully to find information from a single tree crown (Hay et al., 2005). The most applicable methods are region growing and valley following (Erikson & Olofsson, 2005). The first method is based on rule set and region grows start from a seed point to find the boundary of segments. The second method is based on pixels with a lower grey level values than surrounded pixels to find the boundary of objects. Yu et al. (2006) developed a comprehensive vegetation inventory in northern California and demonstrated that the OOA approach is better than pixel based classification. Maier et al. (2008) incorporated LiDAR derived canopy surface models to segment individual trees using multi-resolution segmentation. He has proved that multi-resolution segmentation can be a straight forward method to simplify complex canopy surface with results of 82% of correct trees identified.

3. MATERIALS AND METHODS

This chapter describes the methods followed to address the research questions. The objective of this project is to determine which of the two data sources (LiDAR data and GeoEye images) can provide more accurate information for biomass/carbon estimations.

In order to evaluate the two sensors, the flow of the analysis is presented in Figure 4. It is observed that the method was divided in three main steps defined by the sources of data: GeoEye images, field data and LiDAR data.

The GeoEye Panchromatic image (0.5m) was pan-sharpened with GeoEye Multispectral bands MSS (2m). The pansharpen technique is used to combine within images from the same sensor or by combination with another kind of sensor to produce imagery with higher spatial, temporal or spectral resolution. For this purpose, the low and high-resolution images must be geometrically registered prior to be pansharpened (Section 3.5). The higher resolution image (PAN) was used as the reference to which the lower resolution image (MSS) was registered (Schowengerdt, 2006).

LiDAR data was pre-processed to calculate the canopy height model (CHM), which is the height of trees that was subtracted by the difference between the Digital surface model and the Digital terrain model (DSM – DTM) (Section 3.4).

Both images were filtered (Section 3.6) to reduce the noise and facilitate the segmentation process based on the crown projection area (CPA) of individual trees (Section 3.8). Then values of segmentation were determined with its accuracy in comparison to the manual delineation of trees derived from field data. For this analysis regression analysis and the root mean square error (RMSE) was used. With these accuracy values the most accurate method for delineation crown canopies of individual trees was determined. Then, for biomass estimation an allometric equation described in section 3.10 which considers Diameter at breast height (DBH), Height of trees and wood specific gravity was used.

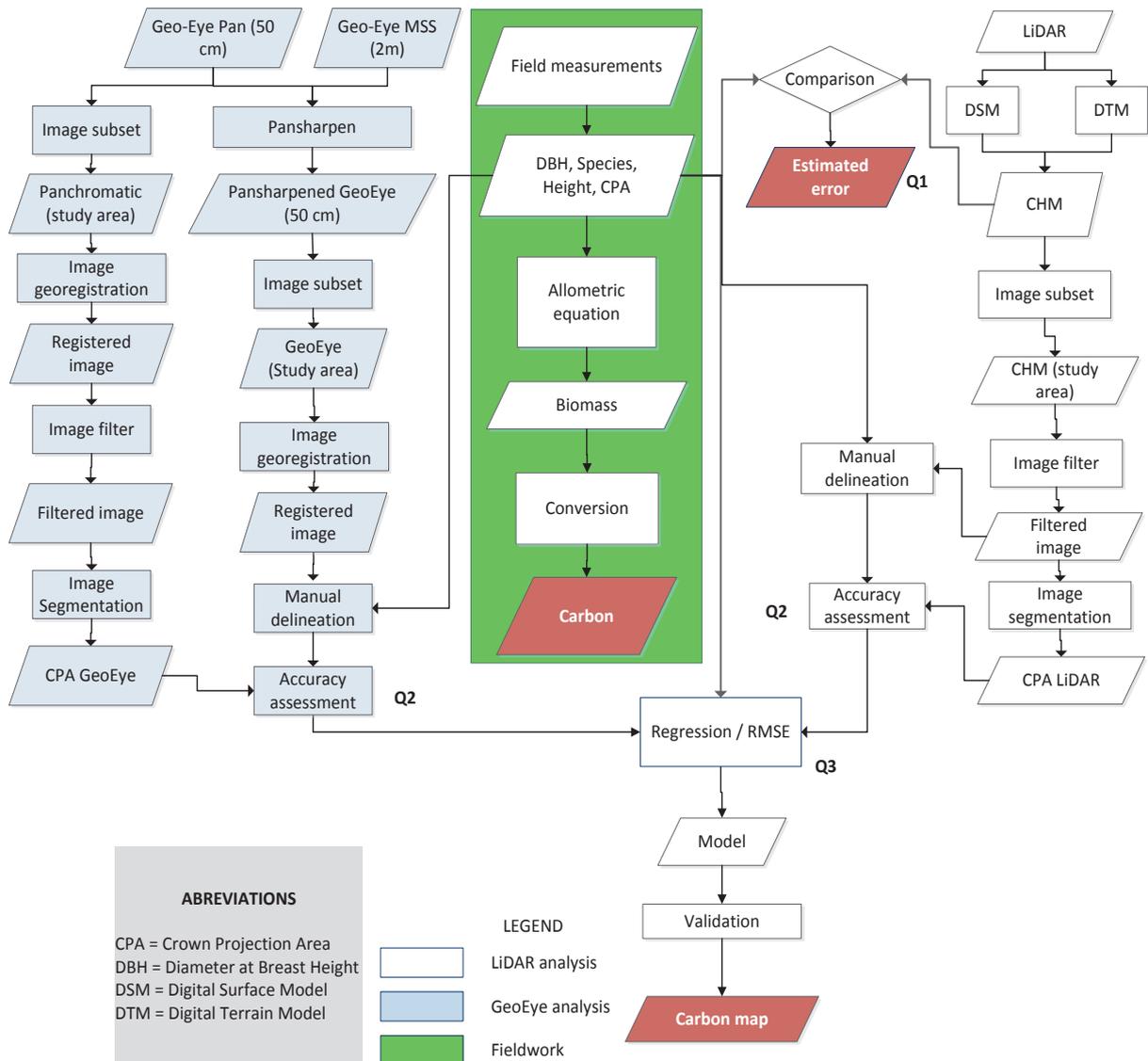


Figure 4 Flowchart for biomass/carbon estimation within GeoEye images and LiDAR data

3.1. Data set

For the study in the Chitwan community forest, two different data were used. LiDAR data was acquired between March and April of 2011 by the Forest Resources Assessment (FRA) project in Nepal and data from GeoEye was acquired in November of 2009 by the ICIMOD project in Nepal. GeoEye images with a ground resolution of 2.0 meters MSS and 0.5 meters in the panchromatic and LiDAR data with an average of 2.4 points/m². Other specifications are presented in Table 1.

Table 1 Dataset characteristics

| Parameters | Sensor | |
|--------------------------------|---|--|
| | LiDAR | GeoEye |
| Costumer | Forest Resource Assessment (FRA) Nepal, Ministry of Forests and Soil Conservation | International Centre for Integrated Mountain Development (ICIMOD), Nepal |
| Projection | UTM | UTM |
| Datum | WGS 84 zone 45N | WGS 84 zone 45N |
| Altitude | 2.2 kilometers | 684 kilometers |
| Aerial platform and orbit type | Helicopter (9N-AIW) | Satellite sun synchronous |
| Band wavelenght | NA | Blue 0.45-0.51 |
| | | Green 0.51-0.58 |
| | | Red 0.655-0.69 |
| | | NIR 0.78-0.92 |
| | | PAN 0.45-0.8 |
| Date flown | 20110316 / 20110328 / 20110401 / 20110402 | NA |
| Flying speed | 80 knots | NA |
| Sensor pulse rate | 52.9 khz | NA |
| Sensor Scan speed | 20.4 lines/second | NA |
| Scan FOW half-angle | 20 degrees | NA |

NA Not Available

3.2. Materials

Additionally to the dataset, materials for the fieldwork are listed in the Table 2 and processes were elaborated in different software's listed in Table 3.

Table 2 List of instruments used for fieldwork

| Instruments | Purpose of usage |
|-------------------------------------|-------------------------------------|
| iPAQ and GPS | Navigation (location of plots) |
| Suunto compass | Orientation (north location) |
| Diameter tape 5 meters | Diameter measurement of trees |
| Measuring tape 30 meters | Length measurement (ratio of plots) |
| Spherical densiometer | Crown cover measurement |
| Slope meter | Slope measurement |
| Laser range finder & Haga altimeter | Tree height measurement |
| Fieldwork datasheet | Field data record |

Table 3 List of software used in the research

| Software | Purpose of usage |
|----------------------------|---------------------------------|
| ArcGIS 10 | GIS analysis |
| Lastools | LiDAR data analysis |
| SCOP++ | |
| Erdas Imagine 2011 | Image processing |
| ENVI 4.8 | |
| eCognition 8.7 | Segmentation and classification |
| R software | Statistical analysis |
| SPSS | |
| Adobe Acrobat Professional | Thesis writing and editing |
| Microsoft Office | |
| End note X5 | |

3.3. Fieldwork

3.3.1. Sampling design

The sampling approach was designed prior to the fieldwork. The stratified random sampling design was applied to collect information in circular plots of 500 m² (radius 12.62 m) (MacDicken, 1997). This sample design helps to ensure that a sample is spread out over the whole study area and divides the population in different homogeneous parts to reduce the error and coefficient of variation. Data from community and previous research projects are provided to help the stratification process. Finally the equation to calculate the number of samples is as follow:

$$n = \frac{t^2 \sum_{j=1}^M P_j s_j^2}{E^2} \quad \text{Equation 1: } n$$

Where,

n = total number of sample units measured for all strata

t = t value derived from significance level and degrees of freedom;

P_j = Proportion of total forest area in jth stratum = N_j/N

s_j² = Variance of X for jth stratum

E = Allowable standard error in units of X (5%) (Husch et al., 2003)

3.3.2. Data collection and analysis

Circular plots with a radius of 12.62 m equivalent to an area of 500 m² were used to measure the forest. In case of plots located on slope, a correction of the radius based on the following equation was applied:

$$re = \frac{r}{\sqrt{\cos\alpha}} \quad \text{Equation 2: } re$$

Where,

r_e = Enlarged radius of the circular plot

r = Plot radius projected onto horizontal plane

α = Angle of slope (van Laar & Akça, 2007)

Each center of plots was geo-referenced and all trees larger than 10 cm DBH located inside of the plot were measured. Biophysical parameters like tree height, crown cover, slope angle and aspect of the plot were collected. Moreover, trees located within the plots and recognizable on the GeoEye image were identified. Finally, through allometric equation the above ground tree biomass was estimated (Section 3.10).

3.4. Canopy Height Model (CHM)

The measurement of LiDAR is a mechanism of distance between the sensor and a target surface. It calculates the time between the emission of the laser pulse and arrival of reflection of that pulse which is known as the return signal. The first return of pulses record only the position of the first object in the path of the laser illumination whereas the last return is a record of the last illumination of the path by the sensor (Figure 5).

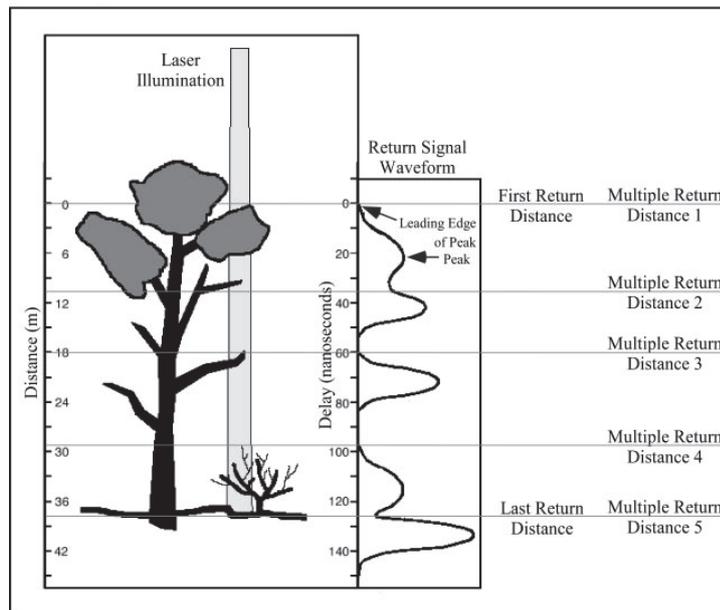


Figure 5 Illustration of the conceptual differences between wave form recording and discrete-return LiDAR devices (Lefsky et al., 2002)

The pulses sent and received by the sensor, can be classified according to the penetration to the target in 9 classes as presented in Table 4. Based on the code classes, extraction of ground and vegetation was performed.

Table 4 Class codes and classification of point clouds

| Class code | Classification type |
|------------|---------------------------|
| 0 | Created, never classified |
| 1 | Unclassified |
| 2 | Ground |
| 3 | Low vegetation |
| 4 | Medium vegetation |
| 5 | High vegetation |
| 6 | Building |
| 7 | Low points (noise) |
| 8 | Model key |
| 9 | Water |

To find the exact locations of individual trees, a Canopy Height Model (CHM) had to be created, by removing the effects of the topography from the raw LIDAR data. The reason for the creation of this model is to accurately identify the actual height of each tree, without the influence of the terrain elevation. In order to create a CHM, a Digital Terrain Model (DTM) and Digital Surface Model (DSM) had to be created and then by subtraction of DSM - DTM was created the CHM (Figure 6). To calculate the CHM different commands were used, and steps were developed as follows:

- *Step 1: Generating a DTM (blast2dem tool)*
Command
`v blast2dem -i cloud_points.las -o chitwan_dtm.tif -v -step 0.5 -keep_class 2`
- *Step 2: Generating a DSM (lasgrid tool)*
Command
`v lasgrid -i cloud_points.las -o chitwan_dsm.tif -first_only -highest -step 0.5 -fill 5 -mem 2000`
- *Step3: Generate Canopy Height Model (CHM)*
Difference between DSM and DTM.

The commands for the first two steps were implemented in the command prompt to generate the rasters, whereas the third step was implemented in the raster calculator of the ArcGis software. Finally, a CHM with 0.5 m spatial resolution was computed which contains pixel values of the height of trees.

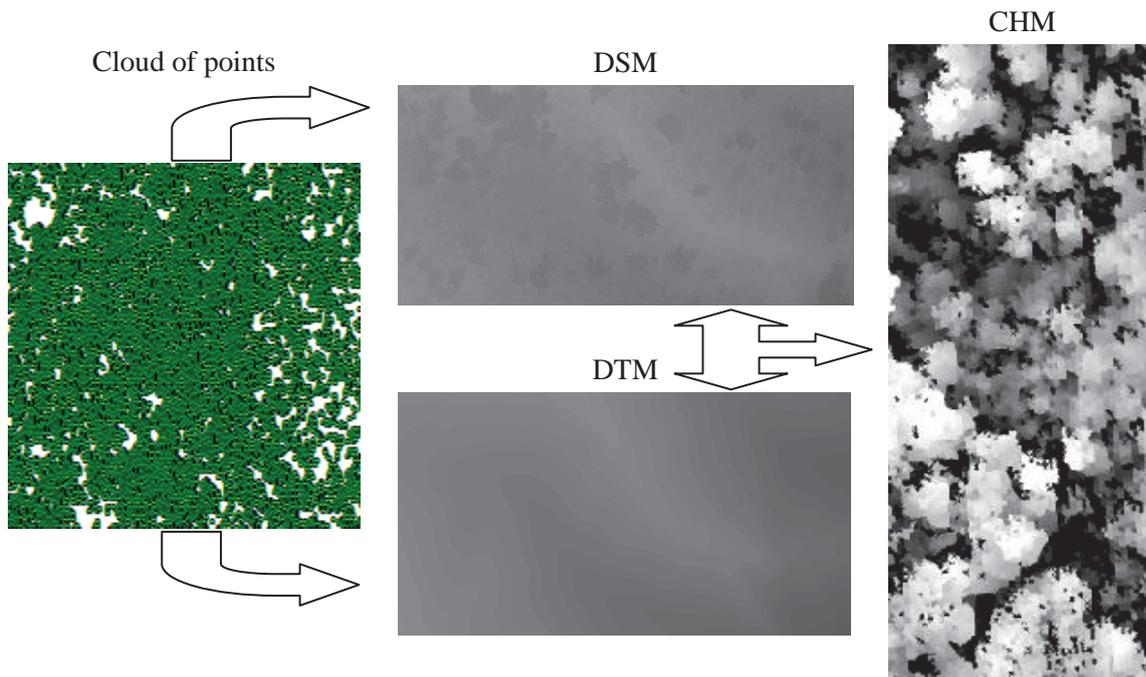


Figure 6 Process to derive Canopy Height Model (CHM) using cloud of pints from LiDAR data

3.4.1. Correlation Analysis

The height of trees collected in the field and the height of trees derived with local maxima from LiDAR data were fitted into a regression model to determine the relationship, the kind of correlation and differences in terms of significance. This analysis is to find out whether it is possible to predict with an accepted accuracy, the height of trees using LiDAR data. The regression model was estimated using excel and R software.

3.5. Image Geo-registration

To geo-register the PAN and Multispectral GeoEye images, an image-to-image transformation was performed using the Erdas Imagine 2011. For the geo-registration root mean square error (RMSE) of 0.36 m was obtained. If the total RMSE value was larger than 0.5 m, the work was repeated until an RMSE value less than 0.5 m was obtained. Then, after correction of the images MS and PAN images were pansharpened to finally obtain GeoEye images with 0.5 m spatial resolution.

3.6. Image filtering

The filter is applied to reduce the noise that may cause over segmentation when individual trees are delineated (Kim & et al, 2010). In this case Gaussian filtering was applied to decrease such unnecessary information on the CHM image and to reduce the intensity variation on LiDAR data using Definiens software. The Gaussian filtering is represented in the following equation:

$$g(x, y) = \frac{1}{2\pi\sigma^2} e^{-\frac{x^2+y^2}{2\sigma^2}} \quad \text{Equation 3: } g$$

Where, x is the distance from the origin in the horizontal axis, y is the distance from the origin in the vertical axis, and σ is the standard deviation which determines the degree of the intensity and emphasizes the cell value in the filtering window (Dralle & Rudemo, 1996), in this case a kernel size of 3 was used.

On GeoEye images a low pass filter was applied to smoothen the appearance on the image with a kernel size of 3. This method is a multiplication of each coefficient in the kernel by the corresponding digital number (DN) in the original image and adding all the resulting products (Avgerinos, 1998).

3.7. Manual delineation of trees

After fieldwork, manual delineation of those recognized trees in the field were identified on the CHM image and also on the GeoEye image. The delineation was performed to validate the automatic delineation of individual trees derived from both images using Definiens software. Though 344 trees were recognized in the field only 144 trees were delineated on LiDAR CHM and GeoEye image.

3.8. Image Segmentation using OOA

Methodologically, this technique is based on three main steps: the first step is partition of an image into a series of non-overlapping, meaningful and homogeneous objects; the second step is the feature set construction and optimization to remove unrelated features, while retaining the useful information to identify the target objects; and the final step is to classify the objects into categories. The method employed for the segmentation of the images, is conducted mainly by the Scale Parameter (SP) which determines the object sizes by measuring the degree of heterogeneity within an image-object. The entire segmentation process was performed with the Definiens software.

3.8.1. Scale Parameter/Setting using ESP

The estimation of the scale parameter (ESP) is based on two main values to build the heterogeneity within an image: Local Variance (LV) and Rate of Change (ROC). The ESP tool generates iterative image objects at different scales to calculate the LV values for each scale. Then thresholds are shown in the ROC of LV (ROC-LV) curve, which indicate the scale were the image can be segmented with more precise values (Dragut, Tiede, & Levick, 2010). Figure 7 shows ESP tool for estimation of scale parameter.

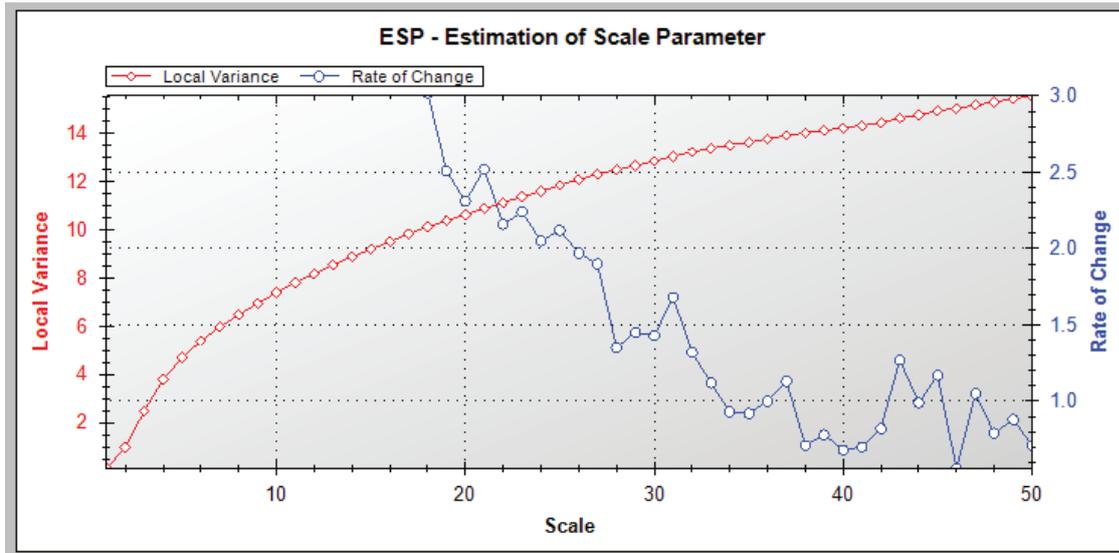


Figure 7 Tool for estimation of Scale Parameter

Initially only the LV was used to define the best thresholds of an image, but with the difficulty to estimate the most appropriate scale parameter in a line which is almost constant, the ROC value was introduced, which is more understandable. As shown on the figure above, the meaningful values are on the levels of 21, 31, 37, 43 and 47 because they are the peaks of the ROC curve. Following this method, scale parameter from CHM of LiDAR data and GeoEye images were estimated. This method helped in the segmentation process to produce only minor modifications in the structure of the objects.

3.8.2. Multi-resolution Segmentation

This segmentation method is a region based algorithm which applies an optimization procedure and minimizes the average heterogeneity of image objects for a given resolution maximizing the homogeneity within an object. The procedure for multi-resolution segmentation can be divided in four main steps: (1) Segments start from single pixels as a starting of objects to merge pixels in series of loops until objects are homogeneous; (2) The starting of pixels looks for its best-fitting neighbour and assembles/merges the neighbour in it; (3) when the best fitting is not mutual, then the best candidate image object becomes the new image object and finds its best fitting partner; (4) Whether the best fitting is mutual, image objects are merged in each loop to proceed with another image object (Definiens, 2007; Rejaur Rahman & Saha, 2008). After the delineation of individual trees, some trees were observed not properly delineated. Then, to solve this problem watershed transformation was used to correct them.

3.8.3. Watershed transformation

Watershed transformation is an algorithm to separate clusters into individual trees. Refine segments on an image considering the image to be processed as topographic surface. Individual objects are split when individual catchment basins touch each other (watersheds) building dams to avoid merging

polygons from different catchments defining those dams as segmentation results (Figure 8) (Derivaux et al., 2010).

Using this algorithm, big crowns and clusters of crowns were split into individual tree crowns. As a parameter in the watershed transformation, a length factor of 8 pixels was decided to use. This size was defined because in the field trees with 4 meters (8 pixels) were found. The outputs of this process are segments with a minimum size of tree crowns of 4 m², but the shapes might not be regular and some segments might not be properly closed, and as a result is necessary to refine the segments to give an approximation of the shape of trees. To transform it in a shape of trees, morphology operation was used.

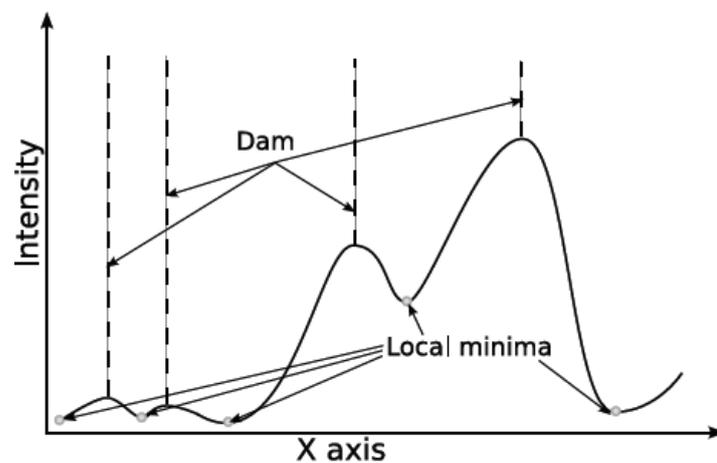


Figure 8 Illustration of the watershed segmentation principle (Derivaux et al., 2010)

3.8.4. Morphology

Morphology is an algorithm to refine segments based on mathematical morphology to smooth the boundaries of objects. The method is based on two basic operations: (1) remove pixels from an image object that are irregular of shape; (2) adds surrounding pixels to an image object to fill small holes inside the segmented area (Definiens, 2007). Modifications of segmented polygons are presented as an example in Figure 9.

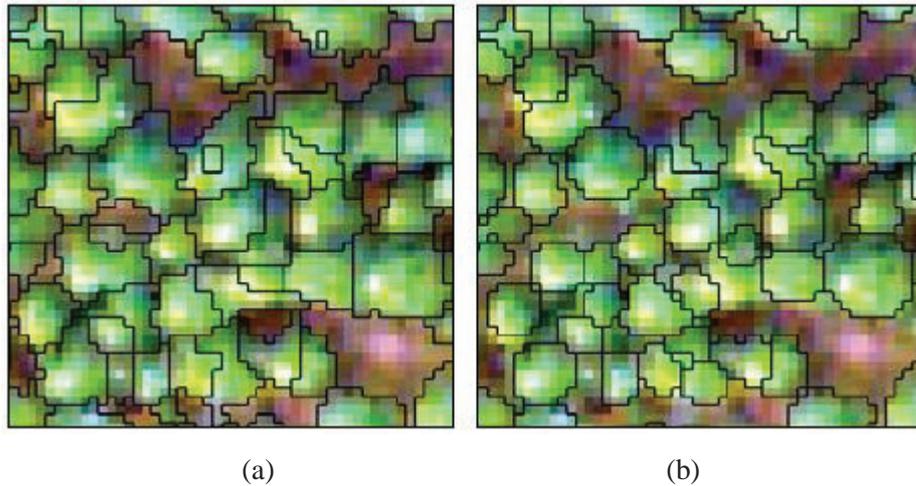


Figure 9 Image object after segmentation (a) Image object after morphology algorithm (b) (Definiens, 2007)

3.8.5. Removal of undesired objects

The undesired objects were removed after the morphological operation. The tiny objects with an area less than 8 pixels (4 meters) were not considered as trees because small objects can be difficult to identify in a dense forest. In the same way, using a filter in Definiens, objects segmented with a high asymmetry were removed, as they are not considered as a normal shape of tree crowns.

3.8.6. Segmentation Validation

Geometric relationships can be determined by a comparison of image positions. As considered by Zhan, et al., (2005), when objects from manual delineation are overlapping by at least 50%, objects are matching, which means that objects take position, size and shape and can be considered as completeness and correctness (Figure 10).

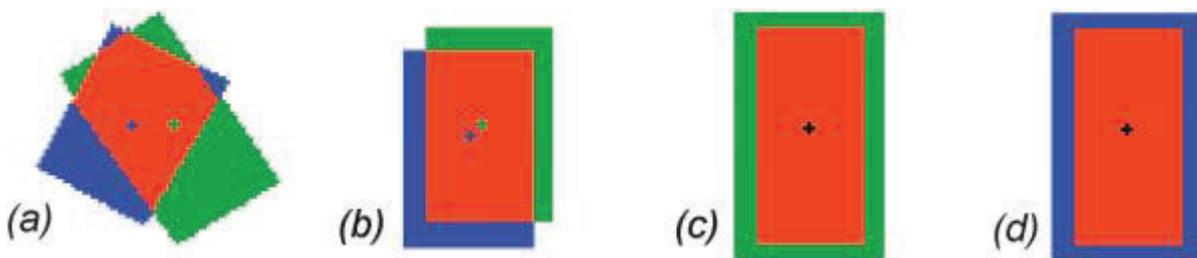


Figure 10 Matched cases of an extracted object (Zhan et al., 2005)

In the figure above, the matching conditions of different objects with automatic segmented objects are shown. (a) a more than 50% match; (b) an extracted reference object matches with the same shape and size but a difference in position; (c) and (d) an extracted reference object matches with the same position but differs in its spatial extent. The reddish part is showing the overlapping of the two objects, while green and blue are differences on the images that are not matching.

Due that measurement the shape similarity between segments and training objects has to match. Then the “D” value was calculated to evaluate the matching of the data. This method is based onto calculated values of OverSegmentation and UnderSegmentation, where the 0 value in both cases define a perfect segmentation, which means that the segments match the training objects exactly (Clinton et al., 2010).

$$D_{ij} = \sqrt{\frac{OverSegmentation_{ij}^2 + UnderSegmentation_{ij}^2}{2}} \quad \text{Equation 4: D}$$

OverSegmentation and Undersegmentation values are calculated as follow:

$$OverSegmentation_{ij} = 1 - \frac{area(x_i \cap y_j)}{Area(X_i)}, y_i \in Y_i^*$$

$$UnderSegmentation_{ij} = 1 - \frac{area(x_i \cap y_j)}{Area(Y_i)}, y_i \in Y_i^*$$

3.9. Model validation

3.9.1. Regression analysis

Regression analysis is to determine the relationship within dependent and independent variables, which means that changes in the independent variable can produce changes in the dependant variable (Husch et al., 2003). Between CPA and DBH a non-linear relationship was established, to evaluate the degree of relation between these two variables and to estimate accuracies to estimate DBH using CPA derived from the segmentation analysis.

For this process, only individual trees matched and recognized in the field and those derived from LiDAR and GeoEye data were used. Out of the 144 trees recognized, only 64 were used for this purpose. Then, the model was validated using a Root mean square error (RMSE) by dividing the data in 70% to develop the model and the remaining values were used for the validation. To calculate the RMSE, the values of DBH observed (field data) and DBH predicted (derived from the model) were compared, based on the following equation.

$$RMSE = \sqrt{\sum \frac{(DBHp - DBHo)^2}{N}} \quad \text{Equation 5: RMSE}$$

Where,

RMSE = Root mean square error

DBHp = Diameter at breast height predicted

DBHo = Diameter at breast height observed

N = Number of observations

3.10. Biomass/Carbon Stock estimation

The most common method to biomass estimation from forest is through allometric equation (Ketterings et al., 2001). Then, the biomass by using a factor value can be converted to carbon stock. Actually different factor values have been defined; in this case biomass is multiplied for a factor of 0.47 to estimate the carbon stock. AGB was estimated using the allometric equation that considers the Height of trees, the Diameter at Breast Height (DBH) and the Wood Specific Gravity (ρ). The equation is presented below.

$$AGB = 0.0509 * \rho D^2 H \quad \text{Equation 6: AGB}$$

Where,

AGB = above-ground tree biomass [kg];

ρ = wood specific gravity [g cm^{-3}];

D = tree diameter at breast height [cm]; and

H = tree height [m].

The above equation was developed by Chave, et al. (2005) on the basis of climate and forest stand types to improve the estimation in tropical and subtropical areas. For wood specific gravity a value of 0.88 was used (ICIMOD et al., 2010) due to the dominance of *Shorea robusta*.

4. RESULTS

4.1. Descriptive statistics

The forest in Chitwan is composed of five community forests. These five community forests were studied and a total number of 1,708 trees were collected from 86 plots. In *Figure 11* the eight most abundant species are presented, representing 88% and in minor presence there are another 31 species representing the remaining 12%.

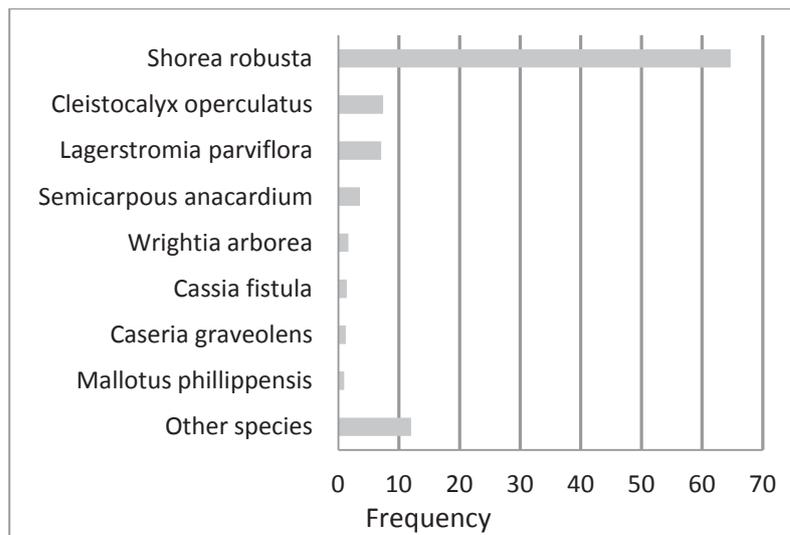


Figure 11 Frequency of forest species

A dominance of *Shorea robusta*, is clearly observed, followed by *Cleistocalix operculatus* and *Lagestromia parviflora* representing 64.7, 7.38 and 7.08 % respectively. The average Height and DBH of trees were analysed and it was found that *Dillenia pentagyna*, *Terminalia alata* and *Lannea caromandelica* are the tallest trees while *Anthocephalus chinensis*, *Albezia procera* and *Lannea caromandelica* are the species with the highest DBH (Table 5).

Table 5 DBH and Height of species observed in the field

| No. | Species | Avg. Height (m) SD=4.36 | Avg. DBH (m) SD=10.2 |
|-----|--------------------------------|-------------------------|----------------------|
| 1 | <i>Dillenia pentagyna</i> | 25.0 | 27.0 |
| 2 | <i>Terminalia alata</i> | 23.8 | 32.0 |
| 3 | <i>Lannea caromandelica</i> | 23.1 | 42.8 |
| 4 | <i>Anthocephalus chinensis</i> | 21.7 | 56.1 |
| 5 | <i>Shorea robusta</i> | 21.5 | 20.7 |
| 6 | <i>Syzygium cumini</i> | 21.4 | 21.4 |
| 7 | <i>Albezia procera</i> | 19.2 | 32.1 |
| 8 | <i>Terminalia chebula</i> | 19.0 | 23.5 |
| 9 | <i>Lagerstromia parviflora</i> | 18.0 | 15.4 |

| | | | |
|----|-----------------------------------|------|------|
| 10 | <i>Spondias pinnata</i> | 16.9 | 28.9 |
| 11 | <i>Schima Wallichii</i> | 16.9 | 25.8 |
| 12 | <i>Careya arborea</i> | 15.8 | 23.5 |
| 13 | <i>Semicarpous anacardium</i> | 14.6 | 20.8 |
| 14 | <i>Albizzia julibrissin</i> | 13.8 | 25.1 |
| 15 | <i>Cleistocalyx operculatus</i> | 13.5 | 15.6 |
| 16 | <i>Pterospermum lanceaefolium</i> | 12.2 | 18.1 |
| 17 | <i>Cassia fistula</i> | 11.6 | 15.3 |

As observed in Table 5, results in standard deviation (SD) indicate that values tend to be close to the mean, showing a low variability from Height values and also for DBH values.

In the same way, carbon was estimated per individual tree, using $AGB=0.0509*\rho D^2H$

Equation 6 and using the species specific parameters presented in Appendix 5. The results are presented in Figure 12. It is observed that the ten highest values are the same species that present the highest values in terms of DBH and Height. The results make sense, because DBH and Height are used as a sole input in the allometric equation. While *Casia fistula*, *Pterospermum lanceaefolium* and *Albizzia julibrissin* present the lowest values with averages of 92, 116 and 199 respectively in terms of carbon and same species present the lowest values in terms of DBH and Height.

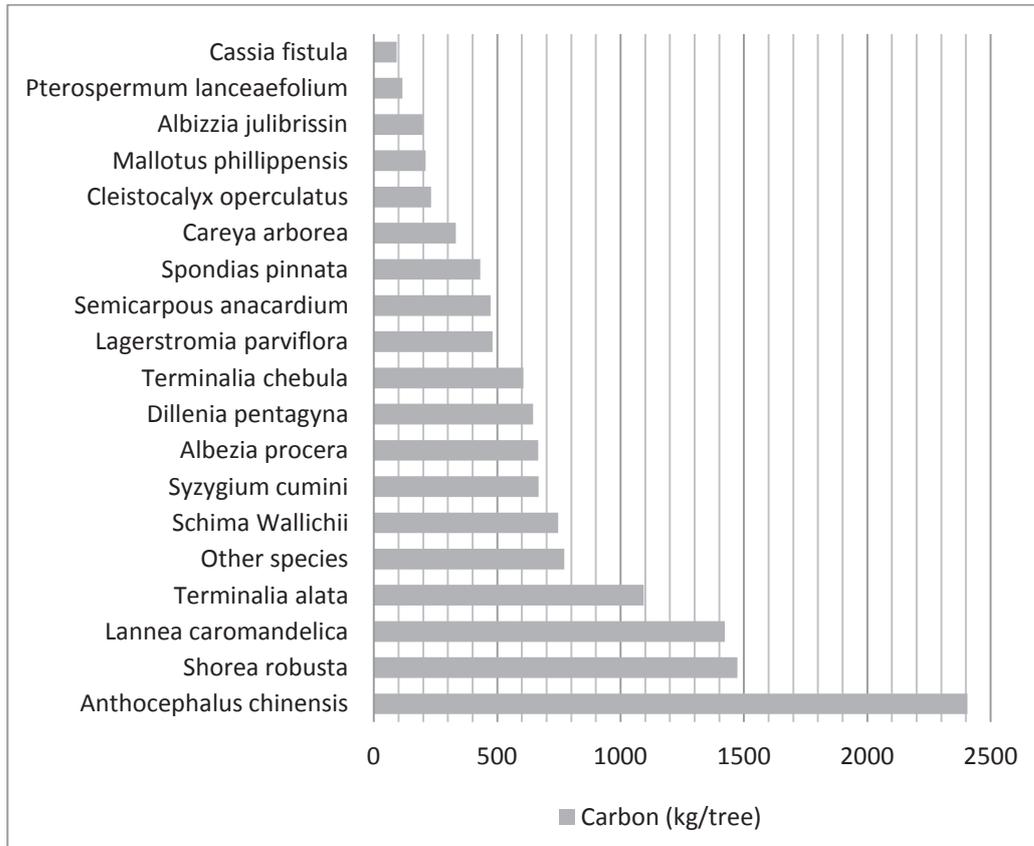


Figure 12 Carbon estimation derived from field data

4.2. Relationship between heights of trees derived from LiDAR CHM and ground data

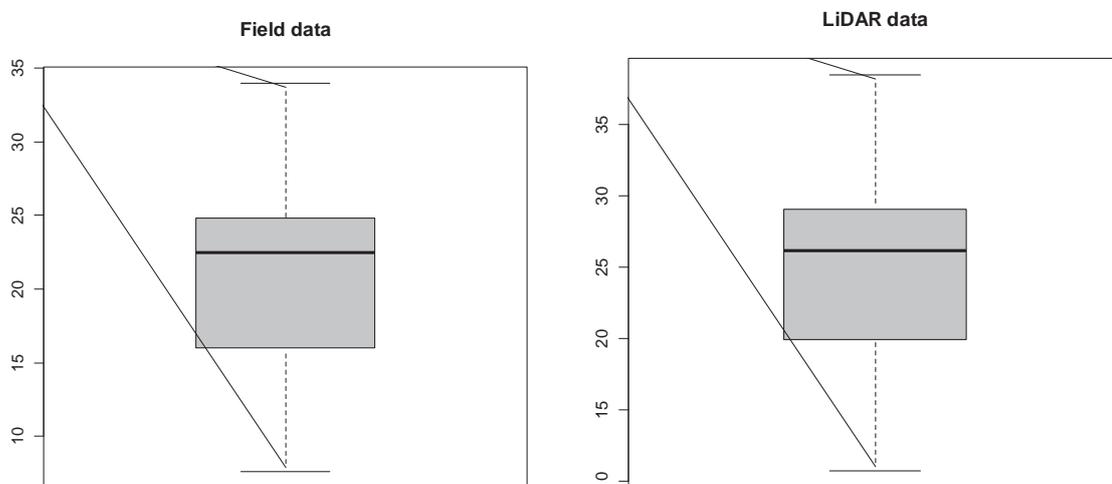
In this regard, the height of trees derived from LiDAR data and height measured in the field were taken and plotted to see the relationship between the two variables. The Shapiro test was calculated to evaluate the normality of the data, presenting p-values of 0.0071 and 0.1695 respectively (Figure 13). With the p values it is observed that LiDAR data does not present the height of trees with normal distribution, contrary to field data. However, the Pearson correlation test was calculated to find out if there is a significant relationship between these two values.

To evaluate if there is significant difference in the height of trees between the two data, t-test was used. First, the F-test was used to evaluate the significance of variances between the two data. Results show that there is no evidence of a significant difference, and then it can be assumed that variances are equal. With this analysis, it shows that the t-test which assumes equality of variances can be used. Result of the t-test at 95% of confidence interval demonstrates that there is a significant relationship between heights of trees derived from LiDAR data when compared to field data, because the t-statistic is greater than t-critical. Results of the t-test and RMSE are presented in Table 6.

Table 6 Results of t-test between tree height values from field and tree height derived from CHM

| Source | Mean | r | R2 | t Stat | P-value | t Crit | RMSE (%) |
|---------------------|-------|------|------|--------|----------|--------|----------|
| <i>Height_LiDAR</i> | 24.45 | 0.81 | 0.65 | 3.06 | 1.36E-03 | 1.98 | 27 |
| <i>Height_field</i> | 20.72 | | | | | | |

Results of the correlation analysis (Figure 14) revealed that there is a positive strong correlation between these two values.



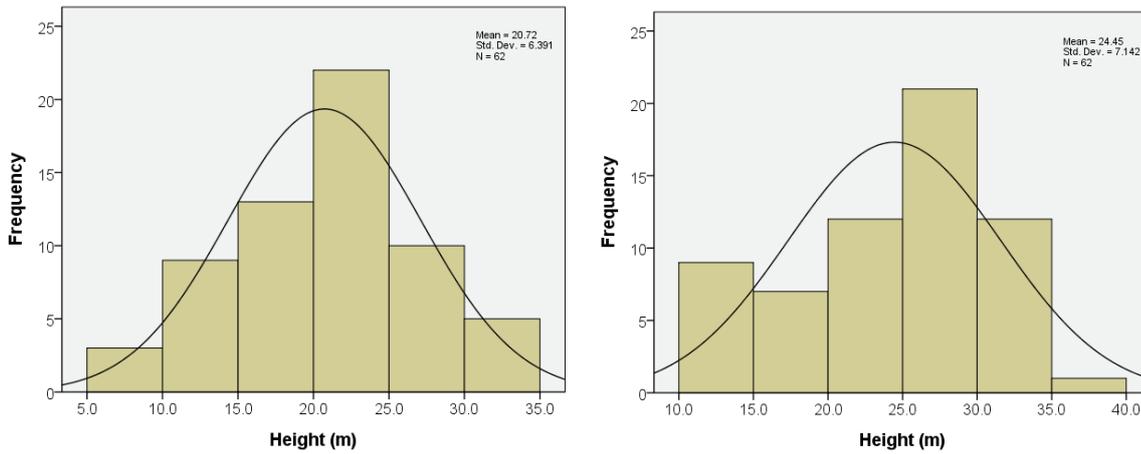


Figure 13 Distribution of tree heights

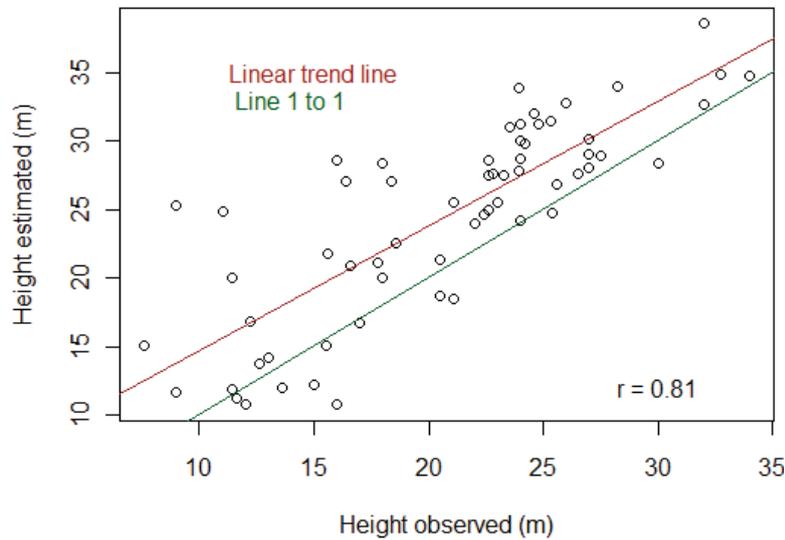


Figure 14 Height values from LiDAR (Height estimated) and height values measured in the field (Height observed) with a fitted line 1 to 1.

The comparison of height values correlated in Figure 14 demonstrates that there is an overestimation of height values derived from LiDAR CHM. It is clearly observed when line 1 to 1 is fitted (green color) falling below the linear trend line of heights from the two datasets.

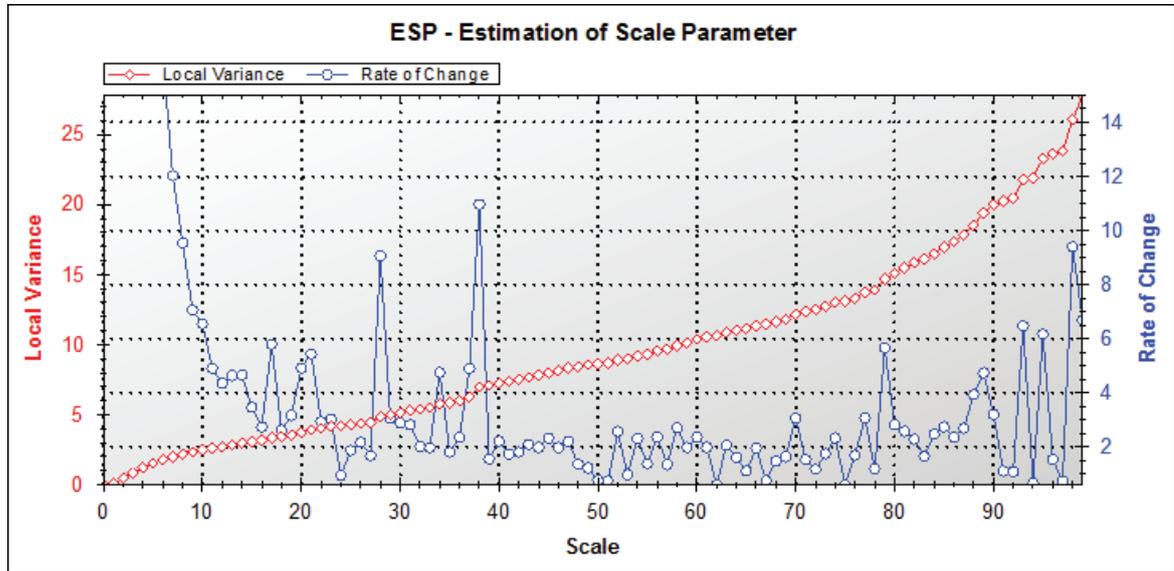
4.3. Image segmentation

Image segmentation was performed on panchromatic GeoEye image and canopy height model (CMH) derived from LiDAR data. Multi-resolution segmentation was carried out to associate pixels into homogeneous areas to separate objects and its results are as follows:

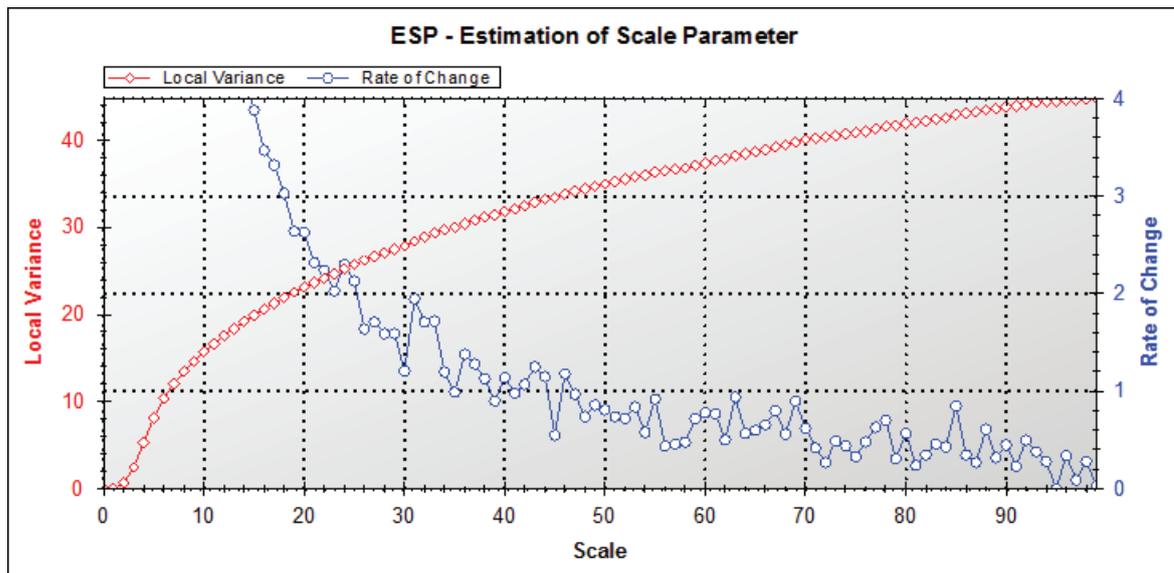
4.3.1. Scale parameter

The result of the scale parameter is presented in Figure 15 and shows the estimation for both images. In the case of the LiDAR image, the peaks of the ROC curve corresponding to scale levels of 17, 21,

28 and 35 were evaluated to segment the image; while on the GeoEye image scale levels of 24, 31 and 36 were evaluated. Finally, in the case of LiDAR CHM, a scale level of 17 was used and in the case of GeoEye, a scale level of 31.



(a)



(b)

Figure 15 Estimation of scale parameter (a) LiDAR CHM, (b) GeoEye image

4.3.2. Multi-resolution segmentation

Multi-resolution segmentation was done using the different scale parameters presented as peaks on the ROC curve above. Then, the goodness of fit (D-value) was calculated for the accuracy of the segmentation; this with the goal to evaluate the existence of over-segmentation or under-segmentation.

Figure 16 shows the output of multi-resolution segmentation.

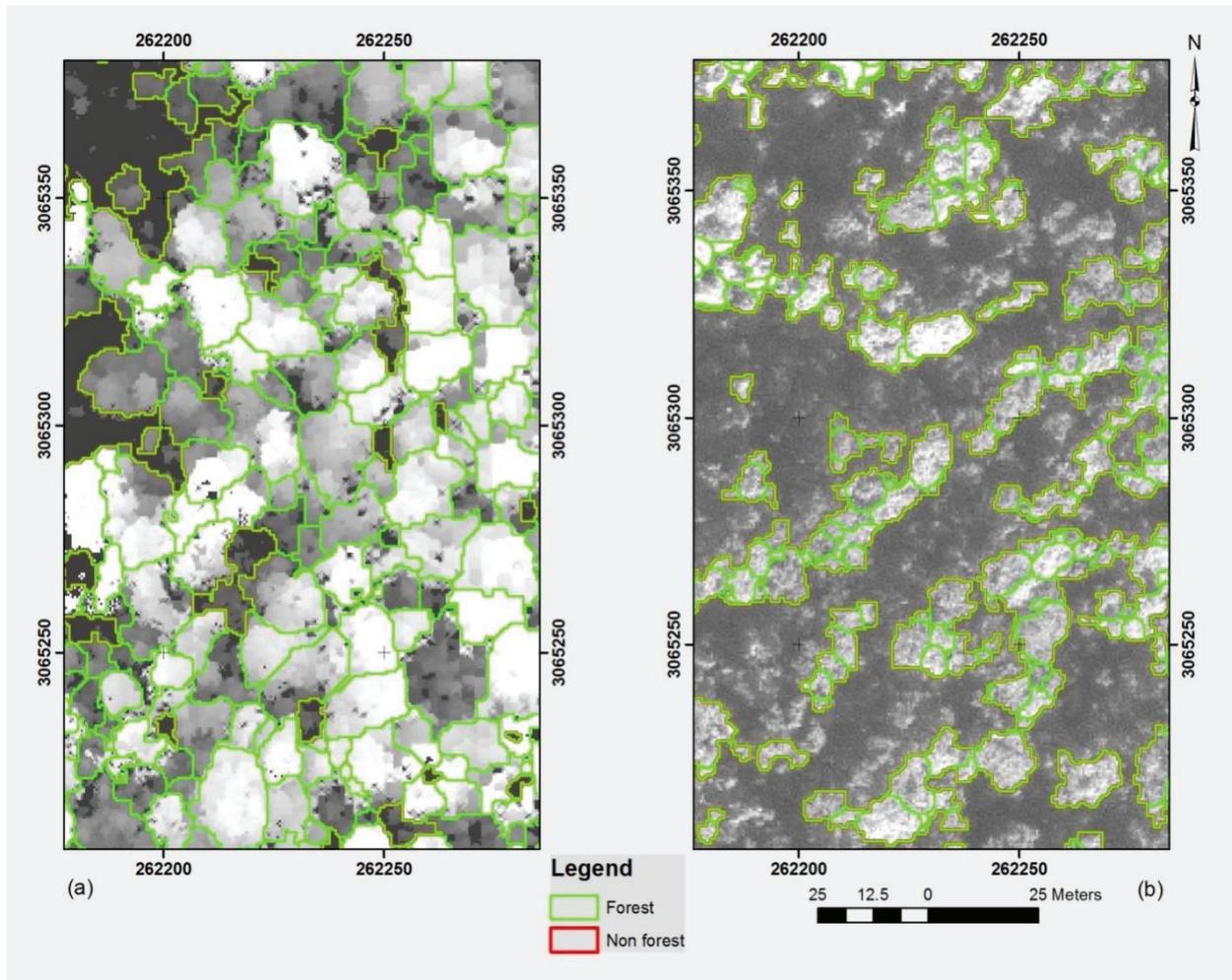


Figure 16 Multi-resolution segmentation and watershed transformation (a) LiDAR CHM, (b) GeoEye

4.3.3. Accuracies of Segmentation

The accuracy assessment of the segmentation was analysed using the goodness of fit of the “D” value. The analysis was in correspondence with the 148 trees that were manually delineated and located on the GeoEye and CHM LiDAR images. For the whole study area, in the LiDAR CHM the over-segmentation value was 0.32 and the under-segmentation was 0.14, while for the GeoEye image values of over-segmentation and under-segmentation were 0.48 and 0.23 respectively. Also r^2 was calculated for both segmentations, results show a greater value in the LiDAR image than Geoeye, with a 0.73 and a 0.59 respectively. Finally, the D values are presented in the Table 7.

Table 7 “D” value from segmentation of LiDAR CHM and GeoEye image

| Sensor | LiDAR | GeoEye |
|--------------------|-------|--------|
| Scale | 17 | 31 |
| Over-segmentation | 0.32 | 0.48 |
| Under-segmentation | 0.13 | 0.23 |
| D-value | 0.24 | 0.38 |

Results presented in Table 7 show overestimations of segments from both images, which is higher on GeoEye image. Remain D-value indicates accuracies of 76% in the case of LiDAR and 64% in the case of GeoEye. As comparison of results Figure 17 shows the frequency of segmented polygons per the whole study area, clearly is observed higher estimation of segments on GeoEye image, accounting higher frequencies for small trees in GeoEye which is reduced on LiDAR CHM image.

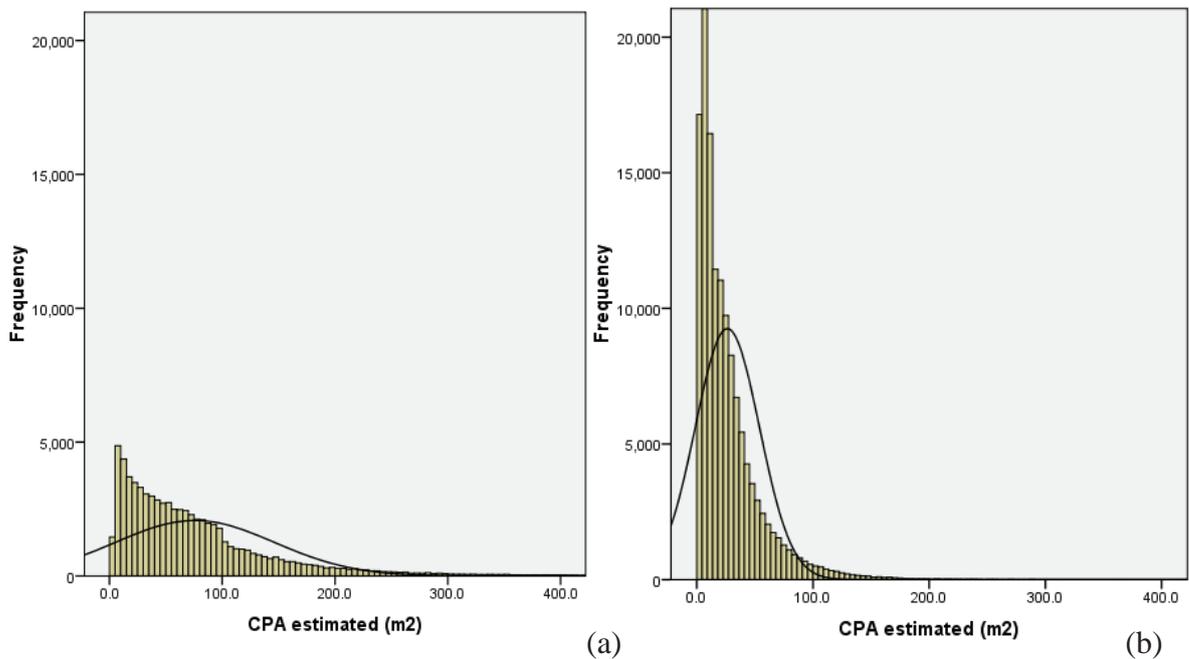


Figure 17 Segmentation derived from LiDAR CHM (a) and GeoEye (b) images

4.4. Relationship between CPA and DBH

The relationship between the crown projection area (CPA) and the diameter at breast height (DBH) was performed to see how accurate DBH can be predicted using CPA values from segments derived from LiDAR CHM and GeoEye images. For this purpose polynomial, logarithmic, linear and power models were evaluated and the result is presented in Table 8. It is observed that polynomial and power models have the lowest value of error. However, all models were developed to derive the regression analysis and RMSE values. Finally only one model was selected for the biomass estimation based mainly on the lowest RMSE value.

The polynomial model was tested parallel to the power model, but present difficulties to predict DBH values for those trees with a wide CPA ($>250 \text{ m}^2$), and then the power model was selected as more appropriate. The first analysis is a comparison of results only within the observed data to see the relationship between CPA and DBH derived from field data (Figure 18).

Table 8 Models for DBH estimations

| Data Source | Model type | Model | R ² | RMSE (cm/tree) | % error | r |
|-------------|-------------|---|----------------|----------------|---------|------|
| Field | Logarithmic | DBH (cm) = 17.746*ln(CPA) - 24.542 | 0.74 | 8.70 | 21.56 | 0.86 |
| | Polinomial | DBH (cm) = -0.001*CPA ² + 0.5337*CPA + 14.374 | 0.79 | 7.45 | 18.48 | 0.90 |
| | Linear | DBH (cm) = 0.2754*CPA + 24.61 | 0.69 | 8.63 | 21.40 | 0.88 |
| | Power | DBH (cm) = 5.6315*CPA ^{0.4956} | 0.81 | 7.51 | 18.63 | 0.89 |
| GeoEye | Logarithmic | DBH (cm) = 22.128*ln(CPA) - 35.786 | 0.72 | 12.48 | 30.94 | 0.71 |
| | Polinomial | DBH (cm) = -0.0015*CPA ² + 0.6206*CPA + 20.133 | 0.64 | 11.85 | 29.37 | 0.71 |
| | Linear | DBH (cm) = 0.2595*CPA + 31.286 | 0.51 | 12.79 | 31.70 | 0.69 |
| | Power | DBH (cm) = 5.1858*CPA ^{0.5555} | 0.64 | 12.10 | 29.99 | 0.72 |
| LiDAR | Logarithmic | DBH (cm) = 26.301*ln(CPA) - 58.416 | 0.84 | 9.70 | 24.04 | 0.84 |
| | Polinomial | DBH (cm) = -0.0018*CPA ² + 0.7534*CPA + 7.7428 | 0.84 | 8.91 | 22.09 | 0.86 |
| | Linear | DBH (cm) = 0.3621*CPA + 21.56 | 0.75 | 9.80 | 24.30 | 0.86 |
| | Power | DBH (cm) = 2.679*CPA ^{0.6835} | 0.79 | 8.54 | 21.17 | 0.86 |

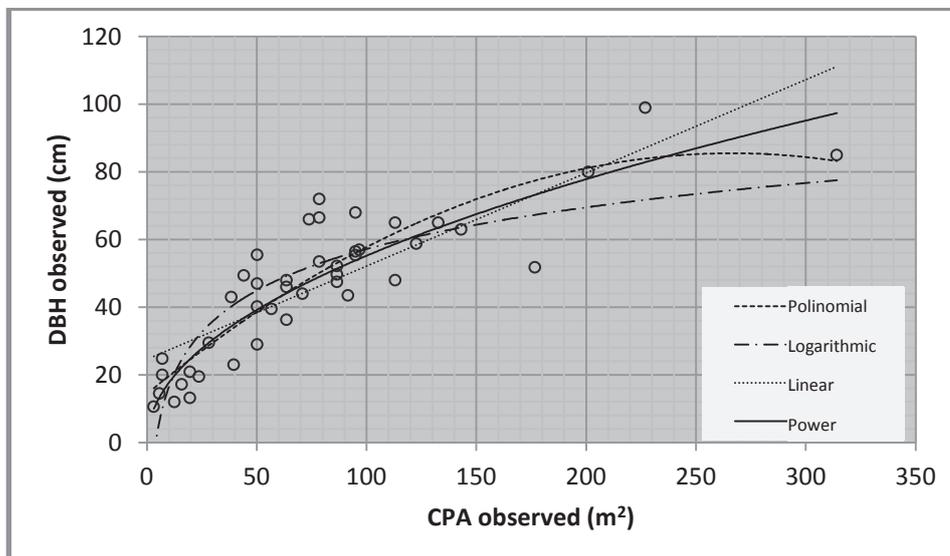


Figure 18 Relationship between CPA and DBH from field data

In the same way, models were calculated and compared for the result of CPA from the GeoEye image and DBH from the field data. The result shows $r^2=0.64$, which means that 64% of DBH can be explained by CPA estimated from GeoEye image (Figure 19) and approximately 30% of error.

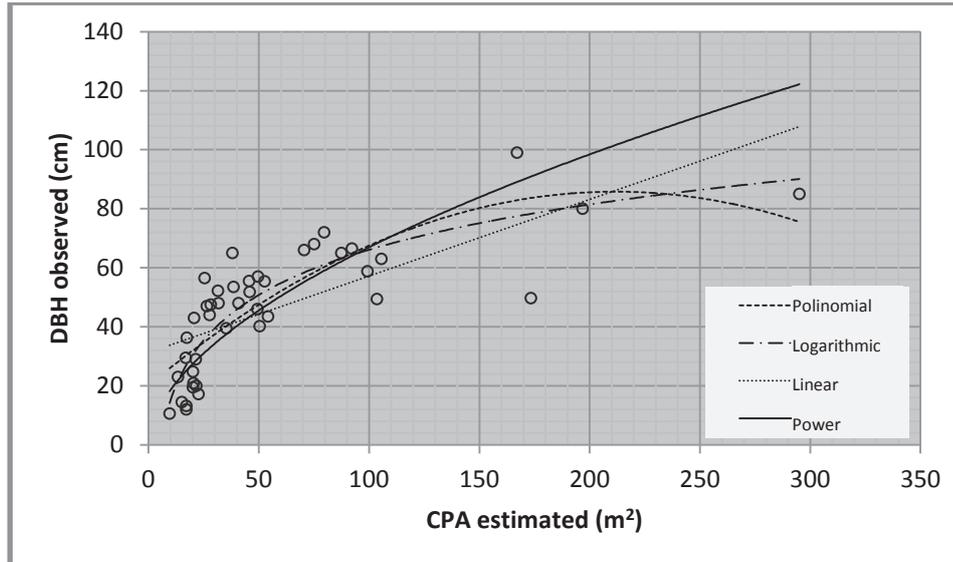


Figure 19 Relationship between DBH from field data (field data) and CPA from GeoEye (CPA estimated)

Afterward, the CPA derived from LiDAR data is contrasted with DBH observed in the field using the same models (Figure 20), result shows that the lowest error is derived by power model with 21% relative RMSE. It also shows that 79% of DBH can be explained by the CPA derived from the LiDAR CHM image.

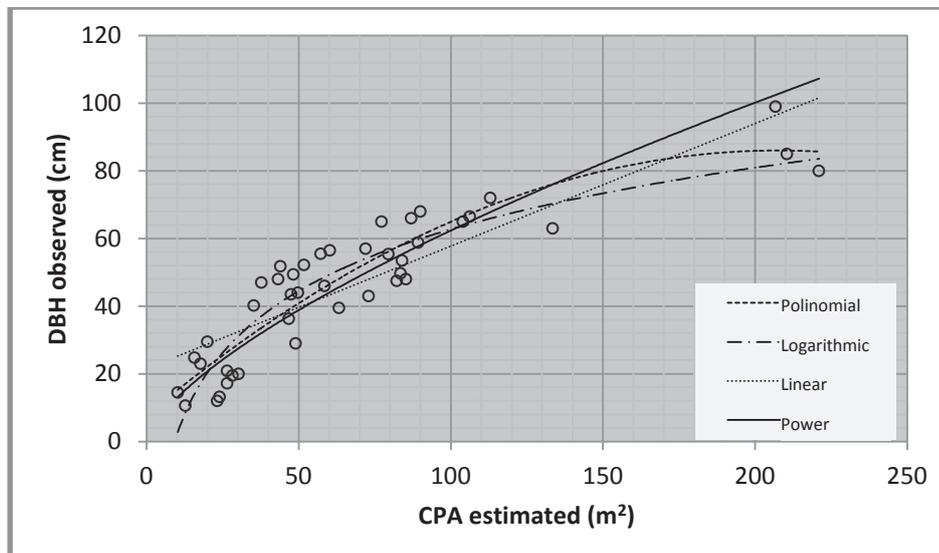


Figure 20 Relationship between DBH from field data (field data) and CPA from LiDAR (CPA estimated)

4.4.1. Model Validation

The model validation was based on the calculation of RMSE values. RMSE values were calculated from the DBH observed in the field and the DBH estimated using the models presented in Table 8. Results demonstrate that using LiDAR data and the power model error is 21% (rmse=8.54) while analysis in GeoEye images is increased to 30% (rmse=12.1). This increasing in error is also producing

a difference in the r^2 of 15% within both images; defining that using LiDAR data can be more accurate for biomass estimation.

In search of the model validation, the DBH predicted from both equations using LiDAR and Geoeye data, were plotted against the DBH measured in the field and calculated a linear regression (Figure 21). The regression line was fitted through the origin for LiDAR and GeoEye derived DBH, results show coefficient of determination of 0.63 and 0.41 while correlation values are 0.86 and 0.72 respectively.

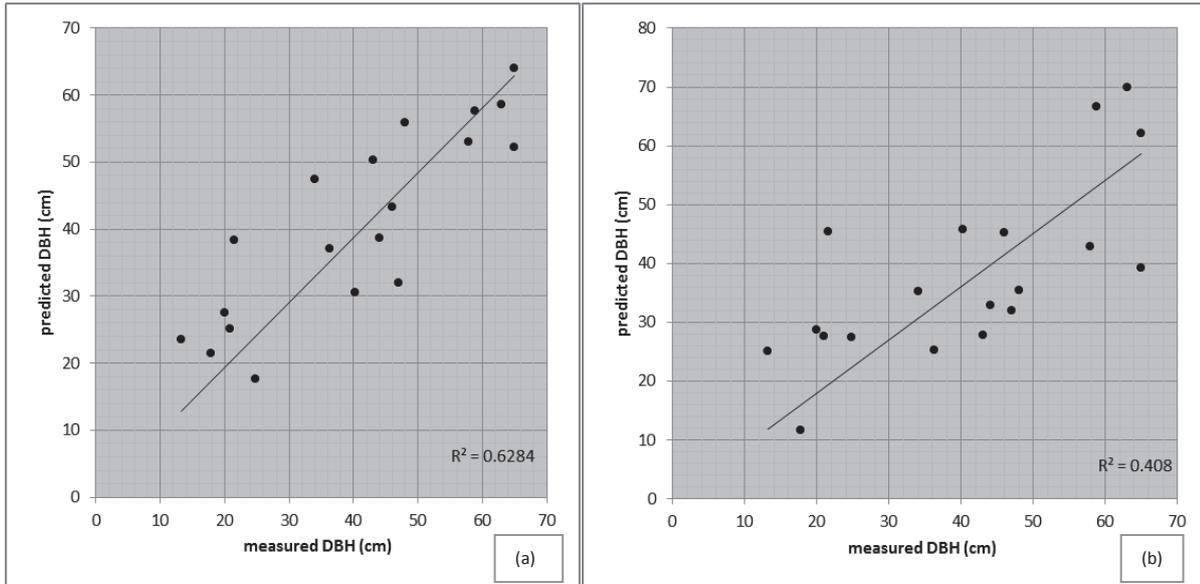


Figure 21 Comparison between the DBH measured in the field and DBH predicted from equations (Table 8) using power models. LiDAR (a) and GeoEye (b)

4.5. Biomass and Carbon Estimation

The biomass estimation was derived using the equation developed by Chave, et al. (2005). Due, the best result of image segmentation and accuracy in DBH estimation was more improved in LiDAR data. This equation considers DBH, height of trees and wood specific gravity. Final equation is presented below.

$$AGTB = 0.0509 * \rho * 2.679 * (CPA)^{0.6835} * H \quad \text{Equation 7: AGTB}$$

Where,

AGTB = above ground tree biomass (kg/tree)

ρ = Wood specific gravity (0.88)

CPA = Crown projection area (m^2) from LiDAR data

H = Height of trees (m)

The final map result for the carbon estimation is presented in Figure 22. The result estimated has a mean value of 1894 kg C/tree equivalent to 181.34 Mg C ha^{-1} .

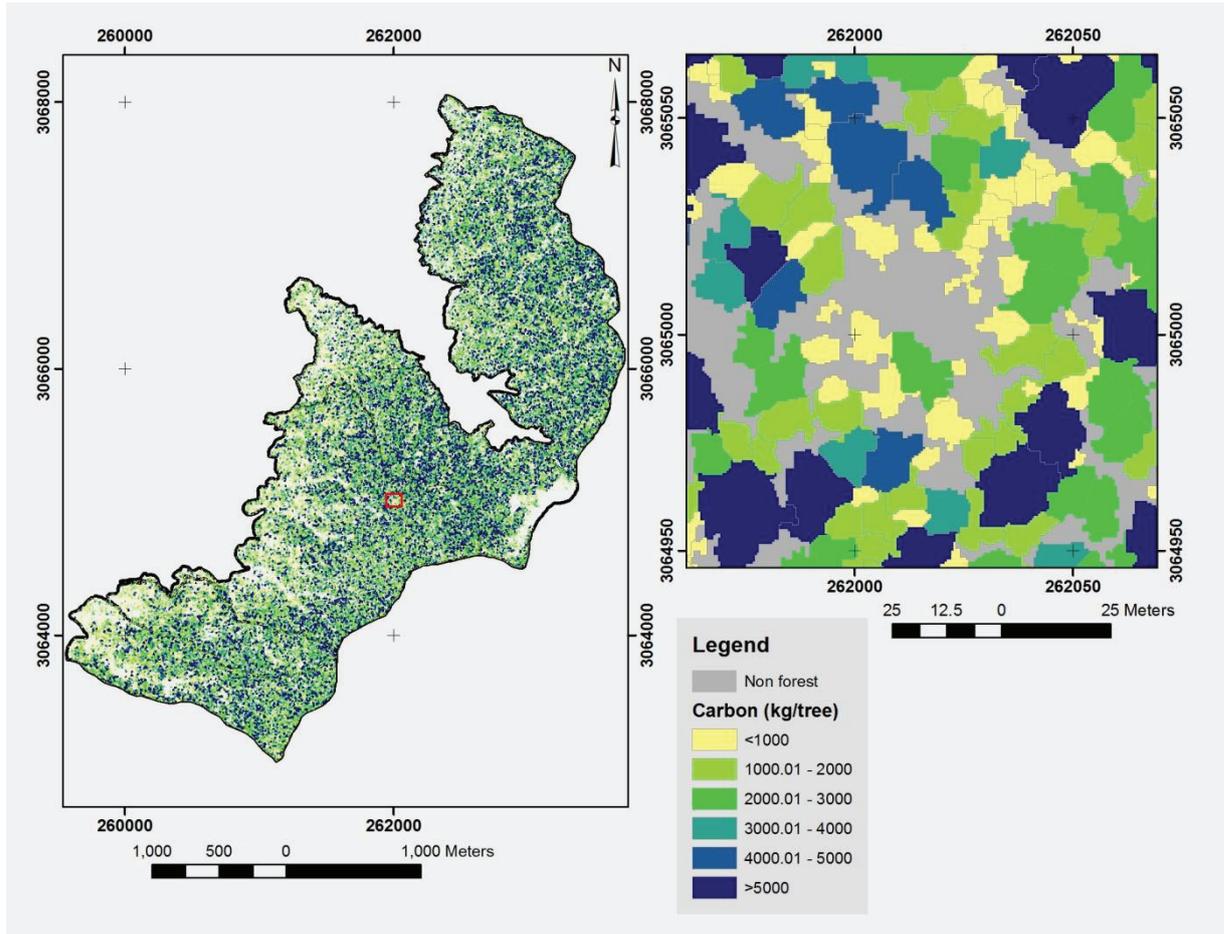


Figure 22 Map of carbon stock in the study area

5. DISCUSSION

5.1. Relationship between height values

Due to the capacity of LiDAR data to estimate the height of trees, it has become widely applicable in forest studies (Naesset, 2002) particularly in forest planning (Naesset, 1997). However, the accuracy still needs to be evaluated.

One of the objectives of this research project was to compare the relationship of tree heights between data collected from the field and heights derived from LiDAR data. The purpose is to evaluate how confident these values are for further analysis related to biomass estimation, which is the main objective of this study.

The error value between LiDAR and field data (27%) is attributed to four main sources: (1) Error derived by the instrument to calculate the height of trees in the field (Laser range finder & Haga altimeter) (2) manipulation of the instrument (3) LiDAR data which was found with noise distributed along the image (4) Difficulty to locate the trees on the LiDAR image, GeoEye image which corresponds to trees on the field. This inaccuracy has caused an overestimation of values on heights derived from LiDAR data (*Figure 23*) compared to the height of trees estimated in the field. The main species with overestimation are *Semicarpous anacardium*, *Pterospermum lanceaefolium* and *Cleistocalix operculatus* that are the most contributing species incrementing the RMSE.

With this analysis it is shown that heights derived from LiDAR data can affect the estimation of biomass when it is used in the allometric equation. Results have shown that only 65 % ($r^2 = 0.65$) of height from trees from the field can be explained by the result of heights derived from LiDAR data, demonstrating large residual error that can affect the estimation of biomass. However, the t-test shows that there is a significant relationship between field and LiDAR data. Trees with the highest residual values are labelled and presented in *Figure 23*.

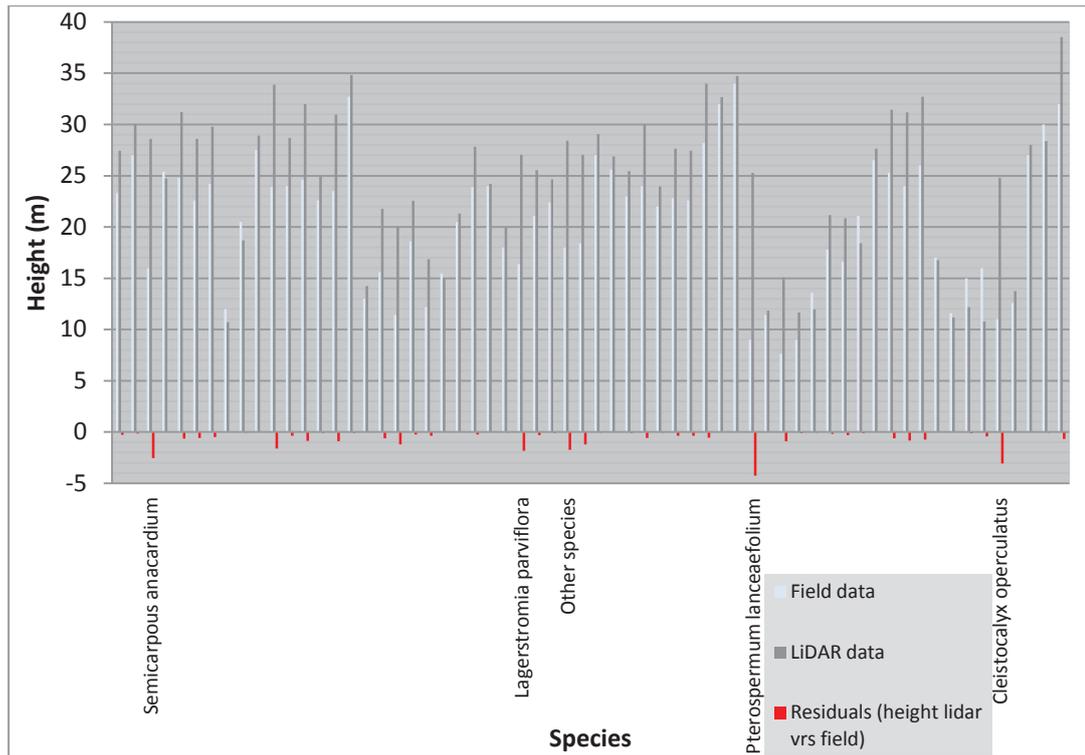


Figure 23 Comparison of Height and mean square errors from trees

5.2. Delineation of tree crowns derived from LiDAR data and GeoEye image

Results of the accuracy obtained are 62% in the case of the GeoEye image and 76% in the case of segments derived from the LiDAR data. Measures applied are based on Zhan, et al (2005), which apply to objects that are matching an object in the reference by at least 50%.

In the case of LiDAR, the result accuracy achieved is similar to Wang, et al., (2004) who used a watershed method to separate individual trees in a forest composed by species of *Picea glauca*, *Pseudotsuga menziesii* and *Abies lasiocarpa*, using a high spatial resolution imagery (0.6 m) and result obtained is 75.6%. As observed, Wang used a different type of images and the species are coniferous, which are easier to segment compared to broadleaves. With these results it can be deduced that the application of LiDAR in coniferous forests should improve better, increasing its potentiality for future studies. Erikson & Olofsson (2005) compared three methods for tree crown detection and found accuracies around 80%, which is also similar to this study. Endo Sawada (2010) presented a method for individual tree crown delineation from point cloud derived from LiDAR data. The study is based on triangulated irregular network (TIN) and found that the delineation of tree crowns on broadleaves was overestimated due to the roughed surface on the same tree. It is mentioned that accuracies can be improved with an estimation method of the tree top position and a good algorithm to determine the search area. Comparing the values of segmented tree crowns in this study with Endo & Sawada, values of overestimation still remain as shown in Table 7.

Differences in accuracies with other studies, are due most of the methods using LiDAR data are applied to Pine species and most of models consider the trees to have a conical shape, which is an ideal situation especially in managed forest or plantations and does not occur in this research project because the entire study area is covered by broadleaved species.

Possible uncertainties are considered also by Koch et al., (2006) in the study detection of individual tree crowns using Airborne LiDAR data. Accuracies obtained was 50% on broadleaves species and 87.3% were found as correctly in the case of soft wood (*Pseudotsuga menziesii*). Differences in accuracy are because the crown of trees in dense broadleaved forest tends to be merged and the segmentation process is more difficult. The author concludes that the algorithms used perform well for coniferous stands. However, in this research the accuracy of segmentation derived from LiDAR data on broadleaves is greater than Koch et al., which is an advantage.

Inaccuracy was also obtained in this study. The main reason is attributed to the forest in the study area that is complex in terms of the geometric shape of trees, the species and age of trees which are considered with a high variation, generating an intermingling situation by the proximity of trees that affect often into merging with surrounding crowns. Hence, the crowns of individual trees become blurred. Another factor is that the terrain is too complex, with slopes found in the field more than 75% especially in the Devidunga community forest. The average slope in the five community forests is 46%.

Concerning to the crown projection area (CPA) derived from both images (LiDAR and GeoEye) compared to crown CPA measured in the field (calculated from crown diameter), results show that the r^2 of GeoEye image is 0.59 and the LiDAR $r^2 = 0.73$. The result derived from GeoEye image is comparable to the result obtained by Brandtberg & Walter (1998) who obtained $r^2 = 0.54$ and mention that a low value in r^2 is due to conditions under the method that is performed with varying altitudes and camera parameters and spatial and spectral resolution. The low values in r^2 obtained in this research can be attributed to the forest type, variation in species and the main age of trees, though in the field big trees (maximum value in crown diameter of 17 meters) were measured and some others are very young with a crown diameter of 2 meters. This high variation does not allow developing a rule-set which can separate and differentiate crowns of trees with a high accuracy.

The final comparison between the segmentation result of the two images (LiDAR and GeoEye) demonstrate that a higher over segmentation value in GeoEye exists than in LiDAR which results in a 14% accuracy of D value differences (Table 7). This implies that, while comparing this difference with

the manual segmentation, the number of crowns derived from GeoEye images are greater than those derived from LiDAR data (See the illustration in Figure 17 and Figure 24).

Another problem found in the GeoEye image is the difficulty to separate low vegetation (below 5 meters) from trees above 5 meters, generating segments on those areas which are non-trees but are considered another type of vegetation (shrubs). While in CHM derived from LiDAR data easily a mask of non-trees was created to those areas with pixel values below 5 meters to avoid overestimation of tree crowns.

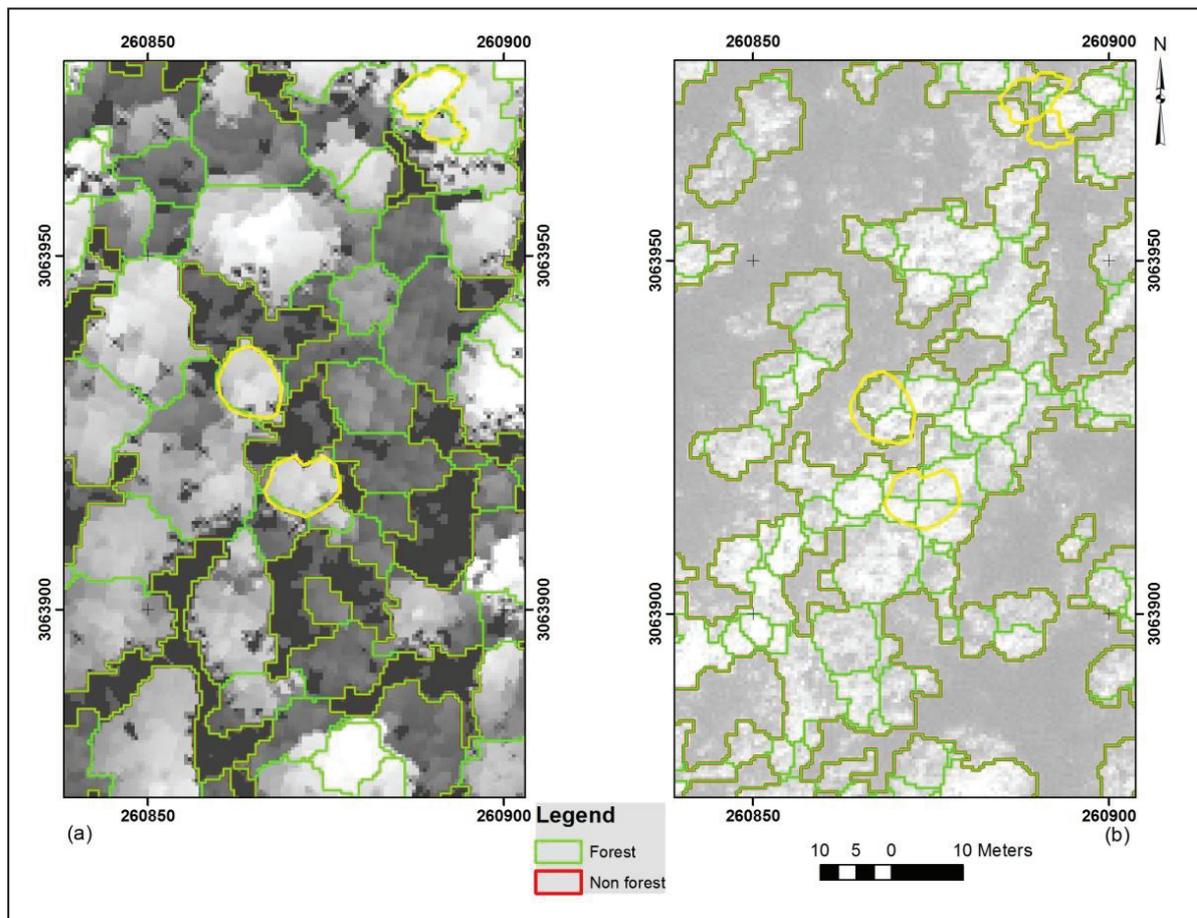


Figure 24 Comparison of segmented trees derived from LiDAR (a) and GeoEye (b)

Based on the analysis above, it can be concluded that results derived from LiDAR data are more accurate when compared to segments derived from GeoEye images. Concluding that segments derived from LiDAR data are more appropriate to DBH estimations.

5.3. Relationship between Crown Projection Area and Diameter at Breast Height

The analysis of the relationship between CPA and DBH was tested with logarithmic, exponential, linear and power models. The RMSE values were calculated as a main factor to decide the most

appropriate model to biomass/carbon estimation, because in this study the objective is to find the model with a lower error for biomass estimation.

Results show that in the case of LiDAR data the highest relative RMSE values were obtained in the linear model with 24.3%, logarithmic model with 24.04% and an exponential model with 22.09% while lowest value was obtained in power model with 21.17%. The results in Table 8 demonstrate that errors increased with data derived from GeoEye images, defining that LiDAR data is more accurate than GeoEye images for DBH predictions.

As further analysis, is observed that power and exponential models were obtained with less error. Then both models were tested to biomass estimations, but at the time of this calculation, the exponential model was unable to extrapolate to predict DBH values to those trees with crown area higher than 250 m². Finally, the power model was defined as most appropriate to perform better than other models. Similarly to this research, Shimano (2000) found that DBH class distribution in a natural broadleaved forest, the power function fits the distribution better than the exponential function.

The correlation value using the training data was calculated between CPA and DBH and result is 0.87. A similar value was found by Shaj et al., (2011) using a linear model while in this study it was found that the power model explains more the relationship between CPA and DBH.

With the comparison of models derived from LiDAR data and GeoEye images, it is defined that LiDAR is most appropriate to biomass estimation because error can be decreased in 8.8% of relative RMSE values and have been shown that there is overestimation in segments derived from GeoEye images when it is compared to segmentation derived from LiDAR data.

5.4. Biomass/Carbon estimation

Different previous studies have attempted to produce biomass regression models in forests located in subtropical areas, including Nepal and more specifically in the Chitwan district. Baral et al., (2011) collected 19 plot data in September - October 2,010 and added 12 plots collected by ICIMOD (2010) in the same forest area of this research study. The author has estimated carbon stock approximately of 70 MgCha⁻¹ using an allometric equation originally developed for East Kalimantan, Indonesia, which considers only the DBH values for biomass estimations. The author determined that 61% of the predicted carbon is explained by the model. At same time and using same method Sha et al., (2011) estimated mean carbon stock of 1394.14 kg/tree for *Shorea robusta*.

Results of Baral et al., (2011) and Sha et al., (2011) differ from this study, despite they were in the same study area. Results of this study estimated a mean average carbon/tree of 1894.08 kg equivalents to 181.34 Mg C ha⁻¹.

Carbon estimation differs with other authors because a different number of plots were measured, probably there is a variation for the different regression models which were affected by the number of plots and it explains the different yield in AGB.

Another factor of variation is that in this study an allometric equation was used which considers a Diameter at breast height (DBH), the Height of trees (H) and wood specific gravity (ρ). Brown et al., (1989) who proposed to include wood specific gravity (ρ) in the allometric equations, suggests that models can lead to an important improvement when this factor is added. Also, Ogawa et al., (1965) found that combining DBH and height (H) was a suitable predictor for AGB instead of using only DBH. Then, in the view that H and ρ factors can improve the biomass estimation rather than using only DBH, it is considered that results in this study are improved in comparison to the previous studies.

5.5. Limitations of the Research

Location of trees

The location of trees in the field were difficult to find due to the fact that the navigation equipment is not precise enough (error was in meters), which carries an error in the location of trees in the images which surely affects the biomass estimation. A Global Position System with high accuracy could be used (like differential GPS) so the location of the trees should be more precise and improved.

Image Orthorectification

Orthorectification of LiDAR data and GeoEye image is important to be used in the fusion of these two images, which could be used to biomass estimation rather than to compare results of the separate images.

LiDAR data

Noise on LiDAR data is a limitation which cannot allow accurate biomass estimation, despite of the fact that a filter was applied, still some pixels with wrong values remain. Hence, LiDAR data is preferred to be corrected previous to any analysis.

6. CONCLUSION AND RECOMMENDATION

6.1. Conclusion

The main objective of this research aims to compare two methods to assess the biomass/carbon estimation using airborne LiDAR data and VHR GeoEye satellite image in a subtropical forest located in Chitwan, Nepal. In this regard, conclusions are based on the research questions as follows:

How strong is the relationship between tree heights measured in the field with height values derived from LiDAR data?

Regarding the relationship between the height values of LiDAR data and field data, a strong relation was found, obtaining an r value of 0.81. Also using the correlation test, a p value of $1.36E-03$ was calculated, which means that there is a strong relationship between these two data and the same result is derived from the t -test where the t -statistic is greater than t -critical. Despite of this, the RMSE value demonstrates that relative error can be increased to 27% which is important to consider in the biomass/carbon estimations.

Does image segmentation from LiDAR data give an improvement in accuracy compared to GeoEye images?

Results show that segmentation using a LiDAR CHM image is more accurate than segmentation using the GeoEye image. LiDAR data yields an accuracy of tree crown delineation of 76% while GeoEye image yields an accuracy of only 62%. Delineating trees in LiDAR data has the advantage that vegetation below five meters was excluded from the analysis, a condition which does not occur in the GeoEye image. This suggests that LiDAR is more accurate to segment tree crowns in dense subtropical forests.

How strong is the relationship between CPA and DBH?

The power model was considered to be the best as a predictor of values of DBH from CPA. Between CPA and DBH, from field data an r^2 of 0.81 was calculated. A similar value was also obtained in the comparison of DBH from the field data and CPA derived from LiDAR data, with $r^2=0.79$, but a weak relationship was found from CPA delineated from GeoEye images with DBH from the field, with $r^2=0.64$. For the validation of the model, the RMSE value was calculated within DBH calculated for those trees identified on LiDAR CHM and Field data. Results show a relative RMSE value of 21.17% for LiDAR data and 29.99% for GeoEye image. This suggests that estimations of DBH derived from CPA of LiDAR data are better than DBH estimated from CPA derived from GeoEye images.

Biomass/carbon estimation in Subtropical forest of Chitwan, Nepal

Since the values of regression and RMSE were higher from segments derived from LiDAR CHM image compared to GeoEye images, then, tree Height, wood specific gravity (ρ) and DBH values were used in the allometric equation for biomass/carbon estimation and the result is $181.34 \text{ Mg C ha}^{-1}$.

6.2. Recommendation

A high accuracy of tree crown delineation in a dense broadleaves forest is difficult to achieve. This study has demonstrated that results in segmentation using LiDAR data are better compared to the segmentation of GeoEye images. However, if these two sensors are fused, it is considered that results can be further improved. This is due to the fact that tree species can be derived from the GeoEye images. This is information that can be well complemented to LiDAR data.

Due to the problem of errors in the process of orthorectification, the combination of LiDAR data with GeoEye image was not possible. It is therefore suggested that the correction of the images to match pixel over pixel can have more accurate estimation.

A general allometric equation for tropical and subtropical areas was used in this study, which introduces error. It is mentioned by the author that using this allometric equation overestimates remain, giving a bias of 0.5-6.5%. Therefore, the development of local allometric equation is required, and if it is possible, per individual species for better carbon stock estimation.

The use of equipment to locate trees (GPS) in the field versus images introduces an error in the estimations. For further analysis, it is suggested to use equipment with a high precision (differential GPS) to ensure the exact location of individual trees.

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APPENDICES

Appendix 1 Sample size for each community forest (CF)

Equation:

$$n = \frac{t^2 \sum_{j=1}^M P_j S_j^2}{E^2}$$

Degrees of freedom = 59

Error allowed = 5%

t = 2.009

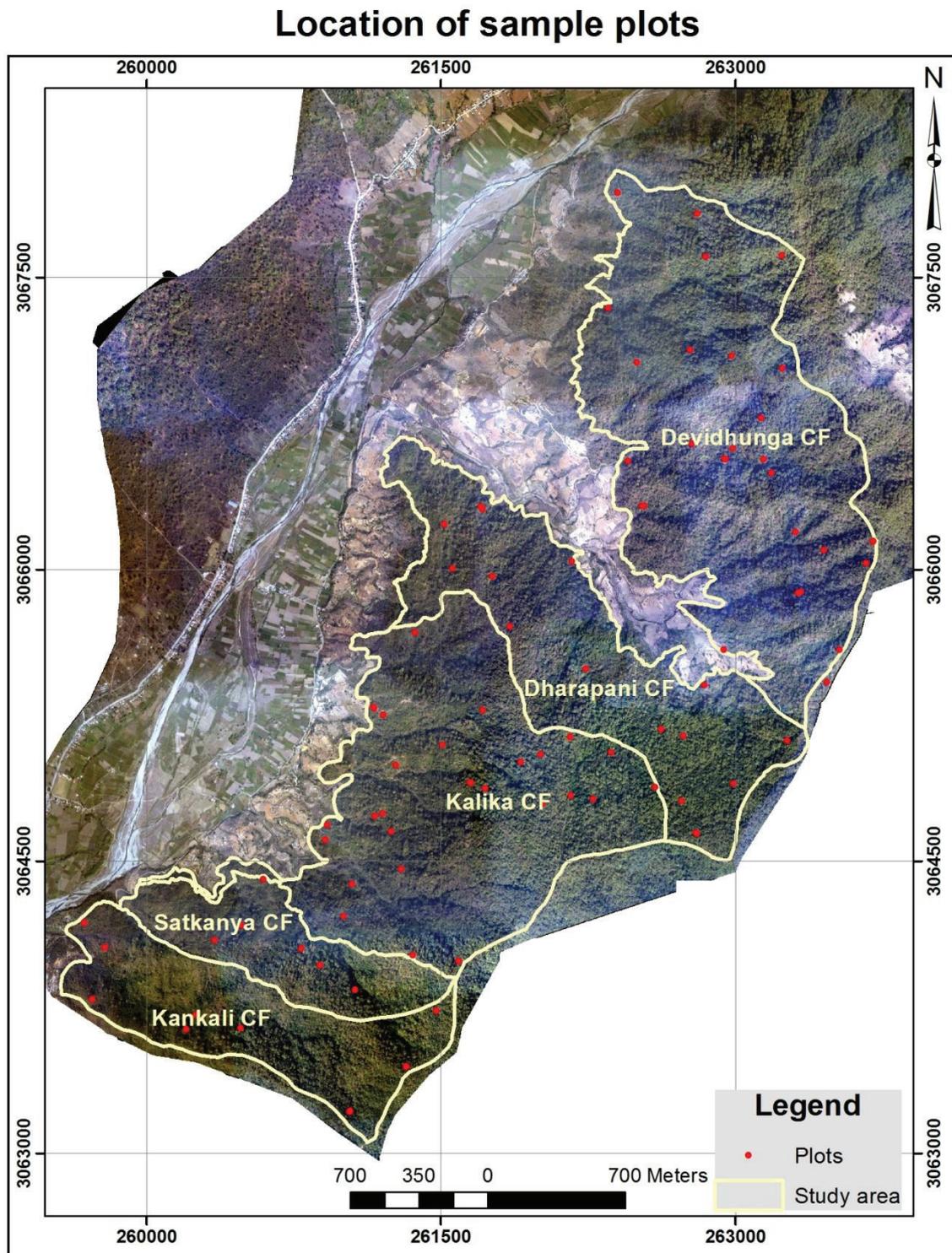
S² = 451

Number of samples per strata:

| Strata | Area | No. Samples |
|---------------|------|-------------|
| Devidhunga CF | 254 | 24 |
| Dharapani CF | 147 | 14 |
| Kalika CF | 214 | 20 |
| Satkanya CF | 58 | 6 |
| Kankali CF | 92 | 9 |
| Total | 765 | 73 |

A total number of 73 plots were planned. However, were collected a total number of 86 plots.

Appendix 2 Locations of total collected sample plots



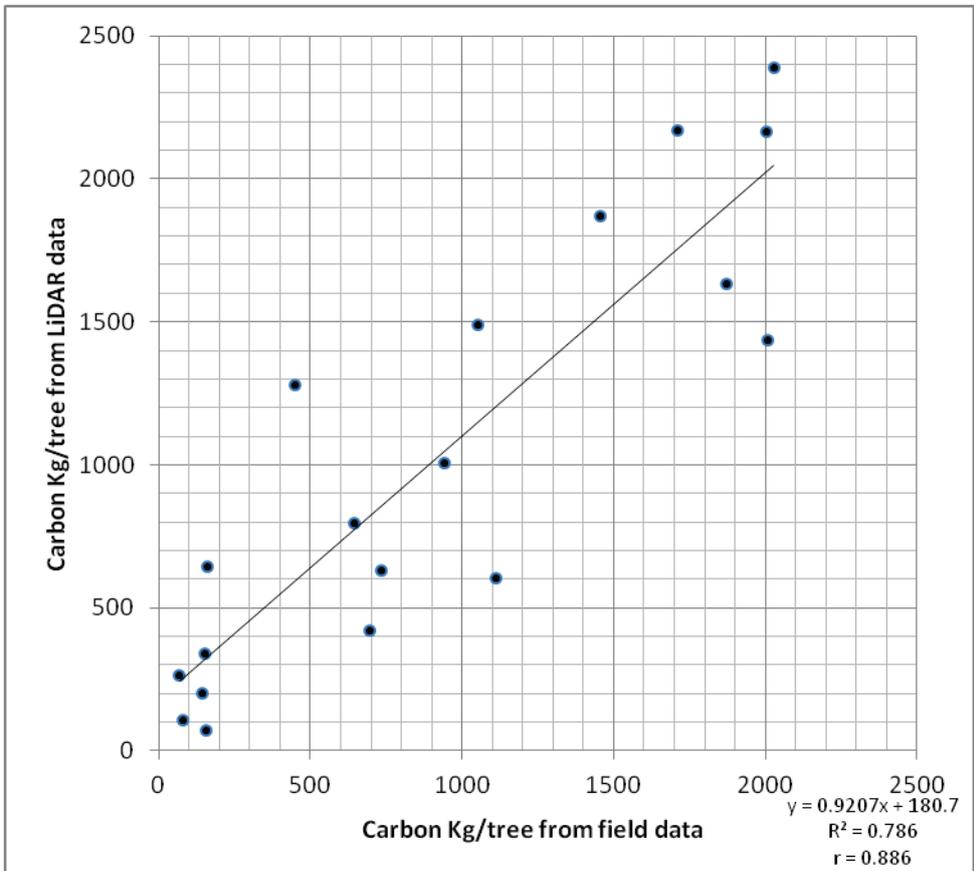
Appendix 3 Recording sheet used in the field

| Name of recorder..... | | | | Date..... | | | | Plot radius size.....12.62m..... | | | |
|-----------------------|-------------|----------|------------|----------------|-----------|-----------------|-----------|----------------------------------|------------|---------------|--------|
| Sampling Plot No. | Coordinates | | | | Elevation | | Slope (%) | | Aspect | | |
| | X | | Y | | | | | | | | |
| Management type: | | | | | | | | | | | |
| Strata Name: | | | | | | | | | | | |
| Forest type: | | | | | | Crown cover (%) | | | | | |
| | | | | | | | | | | | |
| Tree No. | Species | DBH (cm) | Height (m) | Crown diam.(m) | Remark | Tree No. | Species | DBH (cm) | Height (m) | Crown diam(m) | Remark |
| 1 | | | | | | 16 | | | | | |
| 2 | | | | | | 17 | | | | | |
| 3 | | | | | | 18 | | | | | |
| 4 | | | | | | 19 | | | | | |
| 5 | | | | | | 20 | | | | | |
| 6 | | | | | | 21 | | | | | |
| 7 | | | | | | 22 | | | | | |
| 8 | | | | | | 23 | | | | | |
| 9 | | | | | | 24 | | | | | |
| 10 | | | | | | 25 | | | | | |
| 11 | | | | | | 26 | | | | | |
| 12 | | | | | | 27 | | | | | |
| 13 | | | | | | 28 | | | | | |
| 14 | | | | | | 29 | | | | | |
| 15 | | | | | | 30 | | | | | |

Appendix 4 List of local tree species

| <i>Species</i> | <i>Frequency</i> |
|--|------------------|
| <i>Albezia procera</i> | 4 |
| <i>Albizia julibrissin</i> | 4 |
| <i>Anthocephalus chinensis</i> (Lam.) A. Rich. Ex Walp | 9 |
| <i>Bridelia retusa</i> (L.) Spreng | 1 |
| <i>Careya arborea</i> | 16 |
| <i>Caseria graveolens</i> | 21 |
| <i>Cassia fistula</i> | 24 |
| <i>Cissus repens</i> Lam. | 1 |
| <i>Cleistocalyx operculatus</i> | 126 |
| <i>Cornus oblonga</i> (wall.) Sojak | 15 |
| <i>Desmodium oojeinense</i> (Roxb.) Ohashi | 3 |
| <i>Dillenia pentagyna</i> | 4 |
| <i>Ficus racemosa</i> L. | 1 |
| <i>Ficus religiosa</i> L. | 3 |
| <i>Ilex excelsa</i> (Wall.) Hook. | 3 |
| <i>Lagerstromia parviflora</i> | 121 |
| <i>Lannea caromandelica</i> (Houtt.) Merr | 9 |
| <i>Mallotus phillippensis</i> | 17 |
| <i>Manilkara zapota</i> (L.) Van Royen | 7 |
| <i>Myrsine semiserrata</i> | 1 |
| <i>Osbeckia stellata</i> Buch. Ham. Ex D. Don | 5 |
| <i>Pterospermum lanceaefolium</i> Roxb. | 2 |
| <i>Sapium insigne</i> (Royle) Benth. Ex Hook. F | 5 |
| <i>Schima Wallichii</i> | 10 |
| <i>Semicarpous anacardium</i> | 61 |
| <i>Shorea robusta</i> | 1105 |
| <i>Spondias pinnata</i> (L.f.)Kurz | 2 |
| <i>Stereospermum personatum</i> (Hassk.) Chatterjee | 1 |
| <i>Syzygium cumini</i> | 4 |
| <i>Terminalia alata</i> Heyne ex. Roth | 8 |
| <i>Terminalia chebula</i> | 6 |
| <i>Toona ciliata</i> M. Roem | 2 |
| <i>Vernacular name</i> | 79 |
| <i>Wrightia arborea</i> (Dennst.) Mabblerly | 28 |
| <i>Total</i> | 1708 |

Appendix 5 Correlation between carbon stock derived from field and carbon stock estimated from LiDAR data



Appendix 6 Height, DBH, Biomass and carbon of standalone trees from field data

| Scientific name | DBH (cm) | Height (m) | Biomass | Carbon Kg/tree | Scientific name | DBH (cm) | Height (m) | Biomass | Carbon Kg/tree |
|---------------------------------|----------|------------|---------|----------------|---------------------------------|----------|------------|----------|----------------|
| <i>Cleistocalyx operculatus</i> | 19.6 | 7.6 | 130.78 | 61.46 | Vernacular name | 55.5 | 23.9 | 3297.50 | 1549.82 |
| <i>Shorea robusta</i> | 13.5 | 8.2 | 66.94 | 31.46 | <i>Shorea robusta</i> | 50 | 23.3 | 2609.13 | 1226.29 |
| <i>Shorea robusta</i> | 21.4 | 10.6 | 217.44 | 102.20 | <i>Shorea robusta</i> | 68 | 32.7 | 6772.77 | 3183.20 |
| <i>Shorea robusta</i> | 13.8 | 15.5 | 132.22 | 62.14 | <i>Lannea caromandelica</i> | 58.8 | 23.5 | 3639.34 | 1710.49 |
| <i>Shorea robusta</i> | 19.5 | 13.6 | 231.64 | 108.87 | <i>Shorea robusta</i> | 44.5 | 18.6 | 1649.81 | 775.41 |
| <i>Shorea robusta</i> | 17.9 | 11.4 | 163.61 | 76.90 | <i>Shorea robusta</i> | 46 | 23.5 | 2227.33 | 1046.84 |
| <i>Shorea robusta</i> | 81.5 | 31 | 9223.11 | 4334.86 | <i>Shorea robusta</i> | 66 | 19.8 | 3863.26 | 1815.73 |
| <i>Shorea robusta</i> | 75 | 29.5 | 7432.67 | 3493.36 | <i>Shorea robusta</i> | 58 | 20.8 | 3134.15 | 1473.05 |
| <i>Shorea robusta</i> | 77.6 | 24 | 6473.44 | 3042.52 | <i>Shorea robusta</i> | 51.8 | 20.5 | 2463.85 | 1158.01 |
| <i>Shorea robusta</i> | 57.9 | 26.5 | 3979.27 | 1870.26 | <i>Shorea robusta</i> | 57 | 24.6 | 3580.02 | 1682.61 |
| <i>Shorea robusta</i> | 68 | 25.4 | 5260.80 | 2472.58 | <i>Shorea robusta</i> | 91 | 28.8 | 10682.57 | 5020.81 |
| <i>Shorea robusta</i> | 65.8 | 29.6 | 5740.42 | 2698.00 | <i>Shorea robusta</i> | 55 | 24.8 | 3360.30 | 1579.34 |
| <i>Shorea robusta</i> | 72 | 25.3 | 5874.70 | 2761.11 | <i>Shorea robusta</i> | 24.8 | 12 | 330.59 | 155.38 |
| <i>Cleistocalyx operculatus</i> | 32.8 | 11.6 | 558.99 | 262.73 | <i>Cleistocalyx operculatus</i> | 36.8 | 12.8 | 776.44 | 364.93 |
| <i>Cleistocalyx operculatus</i> | 25.4 | 10.9 | 314.99 | 148.04 | <i>Shorea robusta</i> | 85.5 | 19 | 6221.37 | 2924.05 |
| <i>Shorea robusta</i> | 17.2 | 11.4 | 151.06 | 71.00 | <i>Shorea robusta</i> | 65 | 22.6 | 4276.96 | 2010.17 |
| <i>Shorea robusta</i> | 68 | 25.4 | 5260.80 | 2472.58 | <i>Shorea robusta</i> | 91 | 33.5 | 12425.91 | 5840.18 |
| <i>Shorea robusta</i> | 25.5 | 18.6 | 541.74 | 254.62 | <i>Shorea robusta</i> | 45 | 27.8 | 2521.57 | 1185.14 |
| <i>Shorea robusta</i> | 20 | 17.8 | 318.92 | 149.89 | <i>Shorea robusta</i> | 31 | 24.9 | 1071.82 | 503.76 |
| <i>Shorea robusta</i> | 21.5 | 16.6 | 343.70 | 161.54 | Vernacular name | 20 | 9.8 | 175.58 | 82.52 |
| <i>Shorea robusta</i> | 16.6 | 12.4 | 153.05 | 71.93 | <i>Albizia julibrissin</i> | 28.8 | 14.5 | 538.71 | 253.19 |
| <i>Shorea robusta</i> | 19.9 | 16 | 283.81 | 133.39 | <i>Lagerstromia parviflora</i> | 47 | 20 | 1978.91 | 930.09 |
| <i>Shorea robusta</i> | 39.5 | 18.3 | 1278.93 | 601.10 | <i>Shorea robusta</i> | 81.5 | 28.1 | 8360.30 | 3929.34 |
| <i>Shorea robusta</i> | 29 | 21.1 | 794.84 | 373.57 | <i>Terminalia chebula</i> | 27.3 | 13.4 | 447.33 | 210.25 |
| <i>Shorea robusta</i> | 82 | 24.8 | 7469.30 | 3510.57 | <i>Shorea robusta</i> | 58.2 | 20.2 | 3064.77 | 1440.44 |
| Vernacular name | 61.5 | 29.6 | 5014.67 | 2356.90 | <i>Semicarpous anacardium</i> | 26.3 | 13 | 402.77 | 189.30 |
| Vernacular name | 39.2 | 23.8 | 1638.13 | 769.92 | <i>Shorea robusta</i> | 65 | 23.5 | 4447.29 | 2090.22 |
| <i>Shorea robusta</i> | 61.7 | 23.8 | 4058.33 | 1907.42 | <i>Shorea robusta</i> | 59 | 23.2 | 3617.37 | 1700.16 |
| <i>Shorea robusta</i> | 75.5 | 37.2 | 9498.11 | 4464.11 | <i>Shorea robusta</i> | 75 | 29.2 | 7357.09 | 3457.83 |
| <i>Shorea robusta</i> | 65.6 | 30.5 | 5879.06 | 2763.16 | <i>Terminalia alata</i> | 36.3 | 23.3 | 1375.21 | 646.35 |
| <i>Shorea robusta</i> | 73.5 | 30.2 | 7307.72 | 3434.63 | <i>Shorea robusta</i> | 41.3 | 10 | 764.01 | 359.09 |
| <i>Cassia fistula</i> | 21.4 | 11.8 | 242.05 | 113.76 | <i>Shorea robusta</i> | 38 | 17 | 1099.55 | 516.79 |
| <i>Shorea robusta</i> | 14.6 | 10.2 | 97.39 | 45.77 | <i>Semicarpous anacardium</i> | 41 | 16 | 1204.73 | 566.22 |
| Vernacular name | 21.8 | 10.8 | 229.90 | 108.05 | <i>Lagerstromia parviflora</i> | 40 | 16.8 | 1204.01 | 565.88 |
| <i>Shorea robusta</i> | 46.5 | 18.9 | 1830.49 | 860.33 | <i>Shorea robusta</i> | 49.7 | 22.6 | 2500.47 | 1175.22 |
| <i>Shorea robusta</i> | 17.8 | 11.6 | 164.63 | 77.37 | <i>Shorea robusta</i> | 48 | 18 | 1857.61 | 873.08 |
| <i>Shorea robusta</i> | 14.6 | 12.4 | 118.39 | 55.64 | <i>Lagerstromia parviflora</i> | 40 | 24.8 | 1777.35 | 835.35 |
| <i>Shorea robusta</i> | 19 | 12.6 | 203.74 | 95.76 | <i>Shorea robusta</i> | 36.5 | 19.6 | 1169.61 | 549.72 |

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|--------------------------------|------|------|----------|---------|--------------------------------|------|-------|----------|---------|
| <i>Shorea robusta</i> | 11.3 | 9.8 | 56.05 | 26.34 | <i>Shorea robusta</i> | 49 | 21 | 2258.46 | 1061.47 |
| <i>Shorea robusta</i> | 21.1 | 12.1 | 241.30 | 113.41 | <i>Shorea robusta</i> | 36 | 18.8 | 1091.35 | 512.93 |
| <i>Shorea robusta</i> | 14.5 | 13.7 | 129.02 | 60.64 | <i>Terminalia alata</i> | 33.8 | 18.4 | 941.57 | 442.54 |
| <i>Shorea robusta</i> | 18.5 | 13.5 | 206.96 | 97.27 | <i>Shorea robusta</i> | 53.5 | 24.2 | 3102.58 | 1458.21 |
| <i>Terminalia alata</i> | 55.5 | 28.2 | 3890.77 | 1828.66 | <i>Semicarpous anacardium</i> | 49.4 | 16 | 1748.94 | 822.00 |
| <i>Shorea robusta</i> | 74 | 29.8 | 7309.37 | 3435.41 | <i>Lannea caromandelica</i> | 45.5 | 12.2 | 1131.31 | 531.72 |
| <i>Shorea robusta</i> | 50 | 25.7 | 2877.89 | 1352.61 | <i>Terminalia alata</i> | 52.2 | 25.4 | 3100.10 | 1457.05 |
| <i>Shorea robusta</i> | 46 | 21.1 | 1999.86 | 939.93 | <i>Anthocephalus chinensis</i> | 78.2 | 26.2 | 7176.54 | 3372.97 |
| <i>Shorea robusta</i> | 60 | 23.9 | 3853.90 | 1811.33 | <i>Lannea caromandelica</i> | 10.5 | 25.6 | 126.42 | 59.42 |
| <i>Shorea robusta</i> | 55 | 23.4 | 3170.60 | 1490.18 | <i>Shorea robusta</i> | 51 | 33 | 3844.63 | 1806.98 |
| <i>Shorea robusta</i> | 64 | 26.2 | 4806.86 | 2259.23 | <i>Shorea robusta</i> | 47 | 26 | 2572.58 | 1209.11 |
| <i>Shorea robusta</i> | 56.5 | 22.4 | 3202.91 | 1505.37 | <i>Shorea robusta</i> | 49.2 | 33 | 3578.04 | 1681.68 |
| <i>Shorea robusta</i> | 118 | 29.6 | 18461.04 | 8676.69 | <i>Shorea robusta</i> | 51 | 25 | 2912.60 | 1368.92 |
| <i>Shorea robusta</i> | 40.2 | 20.5 | 1483.91 | 697.44 | <i>Shorea robusta</i> | 44 | 27 | 2341.37 | 1100.44 |
| <i>Shorea robusta</i> | 53 | 23.6 | 2969.37 | 1395.60 | <i>Shorea robusta</i> | 35 | 26 | 1426.63 | 670.51 |
| <i>Anthocephalus chinensis</i> | 83.2 | 19.4 | 6015.18 | 2827.14 | <i>Shorea robusta</i> | 72.5 | 18.7 | 4402.69 | 2069.26 |
| <i>Shorea robusta</i> | 65 | 22.8 | 4314.81 | 2027.96 | <i>Anthocephalus chinensis</i> | 51 | 13.7 | 1596.10 | 750.17 |
| <i>Shorea robusta</i> | 54 | 16 | 2089.82 | 982.21 | <i>Shorea robusta</i> | 102 | 25.7 | 11976.61 | 5629.01 |
| <i>Schima Wallichii</i> | 73.2 | 19.2 | 4608.12 | 2165.82 | <i>Vernacular name</i> | 27.7 | 10.35 | 355.71 | 167.19 |
| <i>Shorea robusta</i> | 16.9 | 10.4 | 133.05 | 62.53 | <i>Shorea robusta</i> | 11.9 | 10.8 | 68.50 | 32.20 |
| <i>Shorea robusta</i> | 22.8 | 17.2 | 400.50 | 188.23 | <i>Shorea robusta</i> | 16.5 | 14.6 | 178.04 | 83.68 |
| <i>Shorea robusta</i> | 20.9 | 15.5 | 303.27 | 142.54 | <i>Shorea robusta</i> | 13.7 | 14.4 | 121.06 | 56.90 |
| <i>Shorea robusta</i> | 17.2 | 14.2 | 188.17 | 88.44 | <i>Shorea robusta</i> | 50 | 8 | 895.84 | 421.04 |
| <i>Shorea robusta</i> | 21 | 13.2 | 260.74 | 122.55 | <i>Terminalia chebula</i> | 44 | 24.6 | 2133.25 | 1002.63 |
| <i>Shorea robusta</i> | 17.2 | 12.2 | 161.67 | 75.98 | <i>Shorea robusta</i> | 63 | 24 | 4266.71 | 2005.35 |
| <i>Shorea robusta</i> | 20.4 | 15.6 | 290.79 | 136.67 | <i>Lagerstromia parviflora</i> | 17.8 | 12 | 170.30 | 80.04 |
| <i>Albezia procera</i> | 40.6 | 19.2 | 1417.60 | 666.27 | <i>Lagerstromia parviflora</i> | 25.8 | 16.4 | 488.97 | 229.82 |
| <i>Spondias pinnata</i> | 34.8 | 16.9 | 916.74 | 430.87 | <i>Cassia fistula</i> | 17.1 | 11.4 | 149.31 | 70.18 |
| <i>Lagerstromia parviflora</i> | 41.1 | 18.2 | 1377.07 | 647.22 | <i>Vernacular name</i> | 45 | 16.4 | 1487.54 | 699.14 |
| <i>Shorea robusta</i> | 14.5 | 13 | 122.43 | 57.54 | <i>Lannea caromandelica</i> | 78 | 29.4 | 8011.93 | 3765.61 |
| <i>Shorea robusta</i> | 16.5 | 13.8 | 168.29 | 79.09 | <i>Mallotus phillippensis</i> | 29.5 | 11.4 | 444.37 | 208.86 |
| <i>Shorea robusta</i> | 72 | 8 | 1857.61 | 873.08 | <i>Shorea robusta</i> | 49 | 26.2 | 2817.69 | 1324.32 |
| <i>Shorea robusta</i> | 13.5 | 2.5 | 20.41 | 9.59 | <i>Shorea robusta</i> | 23 | 12.4 | 293.82 | 138.09 |
| <i>Shorea robusta</i> | 13.2 | 18.6 | 145.16 | 68.23 | <i>Shorea robusta</i> | 33.5 | 19.8 | 995.30 | 467.79 |
| <i>Shorea robusta</i> | 66.5 | 22.6 | 4476.64 | 2104.02 | <i>Shorea robusta</i> | 55.4 | 25.6 | 3519.33 | 1654.08 |
| <i>Shorea robusta</i> | 56.7 | 22 | 3168.03 | 1488.97 | <i>Shorea robusta</i> | 28 | 23.6 | 828.76 | 389.52 |
| <i>Shorea robusta</i> | 78.9 | 25 | 6970.99 | 3276.37 | <i>Shorea robusta</i> | 47 | 23.9 | 2364.80 | 1111.46 |
| <i>Shorea robusta</i> | 75 | 23.5 | 5920.94 | 2782.84 | <i>Shorea robusta</i> | 34 | 18.4 | 952.74 | 447.79 |
| <i>Shorea robusta</i> | 55.8 | 25.9 | 3612.17 | 1697.72 | <i>Shorea robusta</i> | 44 | 18 | 1560.91 | 733.63 |
| <i>Shorea robusta</i> | 48.4 | 24.4 | 2560.24 | 1203.31 | <i>Shorea robusta</i> | 50 | 23 | 2575.54 | 1210.50 |

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|-----------------------------------|------|------|---------|---------|-----------------------------------|------|------|----------|---------|
| <i>Shorea robusta</i> | 68.7 | 29 | 6130.73 | 2881.44 | <i>Shorea robusta</i> | 47.5 | 23 | 2324.42 | 1092.48 |
| <i>Shorea robusta</i> | 38.5 | 29.2 | 1938.67 | 911.18 | <i>Shorea robusta</i> | 52 | 24 | 2906.82 | 1366.21 |
| <i>Shorea robusta</i> | 78 | 29.2 | 7957.42 | 3739.99 | <i>Shorea robusta</i> | 43.5 | 31.5 | 2669.87 | 1254.84 |
| <i>Shorea robusta</i> | 57.2 | 27 | 3956.91 | 1859.75 | <i>Shorea robusta</i> | 42 | 23.5 | 1856.81 | 872.70 |
| <i>Lagerstromia parviflora</i> | 14.1 | 17.8 | 158.51 | 74.50 | <i>Shorea robusta</i> | 52 | 33.5 | 4057.44 | 1907.00 |
| <i>Syzygium cumini</i> | 38.5 | 21.4 | 1420.81 | 667.78 | <i>Anthocephalus chinensis</i> | 27.5 | 21 | 711.35 | 334.34 |
| <i>Shorea robusta</i> | 77 | 29.8 | 7914.04 | 3719.60 | <i>Shorea robusta</i> | 63 | 32 | 5688.94 | 2673.80 |
| <i>Shorea robusta</i> | 66 | 20.8 | 4058.37 | 1907.43 | <i>Shorea robusta</i> | 99 | 34 | 14926.22 | 7015.32 |
| <i>Shorea robusta</i> | 81.5 | 20.4 | 6069.40 | 2852.62 | <i>Shorea robusta</i> | 50 | 11 | 1231.78 | 578.94 |
| <i>Shorea robusta</i> | 56.5 | 19.8 | 2831.15 | 1330.64 | <i>Shorea robusta</i> | 36 | 23 | 1335.16 | 627.53 |
| <i>Shorea robusta</i> | 75 | 35 | 8818.43 | 4144.66 | Vernacular name | 34 | 18 | 932.03 | 438.06 |
| <i>Shorea robusta</i> | 59.5 | 15.6 | 2473.77 | 1162.67 | <i>Shorea robusta</i> | 43 | 27 | 2236.15 | 1050.99 |
| <i>Dillenia pentagyna</i> | 35 | 25 | 1371.76 | 644.72 | <i>Shorea robusta</i> | 48 | 30 | 3096.02 | 1455.13 |
| <i>Shorea robusta</i> | 23 | 11 | 260.64 | 122.50 | <i>Shorea robusta</i> | 61 | 28 | 4666.79 | 2193.39 |
| <i>Shorea robusta</i> | 76 | 25.4 | 6571.45 | 3088.58 | <i>Shorea robusta</i> | 66 | 32 | 6243.65 | 2934.51 |
| <i>Lannea caromandelica</i> | 58.8 | 27.6 | 4274.29 | 2008.92 | <i>Shorea robusta</i> | 19.3 | 17 | 283.64 | 133.31 |
| <i>Shorea robusta</i> | 53.5 | 27.6 | 3538.48 | 1663.09 | <i>Cleistocalyx operculatus</i> | 17 | 18 | 233.01 | 109.51 |
| <i>Shorea robusta</i> | 51 | 26 | 3029.10 | 1423.68 | <i>Shorea robusta</i> | 17.8 | 18 | 255.45 | 120.06 |
| <i>Shorea robusta</i> | 43.5 | 27.5 | 2330.84 | 1095.49 | <i>Shorea robusta</i> | 57 | 33 | 4802.46 | 2257.16 |
| <i>Shorea robusta</i> | 36.5 | 20.6 | 1229.29 | 577.77 | <i>Cleistocalyx operculatus</i> | 48 | 11 | 1135.21 | 533.55 |
| <i>Shorea robusta</i> | 24 | 20 | 516.00 | 242.52 | <i>Shorea robusta</i> | 16.5 | 12 | 146.34 | 68.78 |
| <i>Schima Wallichii</i> | 23.1 | 16.2 | 387.20 | 181.99 | <i>Shorea robusta</i> | 17 | 16 | 207.12 | 97.35 |
| <i>Shorea robusta</i> | 40.5 | 18.4 | 1351.85 | 635.37 | <i>Cleistocalyx operculatus</i> | 21.5 | 19 | 393.40 | 184.90 |
| <i>Shorea robusta</i> | 74 | 25.6 | 6279.19 | 2951.22 | <i>Cleistocalyx operculatus</i> | 23 | 17 | 402.81 | 189.32 |
| <i>Schima Wallichii</i> | 34 | 19.2 | 994.17 | 467.26 | <i>Albizia julibrissin</i> | 23 | 13 | 308.03 | 144.78 |
| <i>Shorea robusta</i> | 70 | 28.5 | 6255.20 | 2939.95 | <i>Shorea robusta</i> | 15.5 | 14 | 150.66 | 70.81 |
| <i>Shorea robusta</i> | 37.3 | 25.2 | 1570.43 | 738.10 | <i>Shorea robusta</i> | 17.5 | 15 | 205.76 | 96.71 |
| <i>Shorea robusta</i> | 17.8 | 17.9 | 254.03 | 119.40 | <i>Schima Wallichii</i> | 25 | 13 | 363.94 | 171.05 |
| <i>Lannea caromandelica</i> | 32.8 | 20.4 | 983.06 | 462.04 | <i>Shorea robusta</i> | 99.6 | 33 | 14663.35 | 6891.77 |
| <i>Shorea robusta</i> | 16.3 | 14.1 | 167.80 | 78.87 | <i>Shorea robusta</i> | 96.6 | 28 | 11703.42 | 5500.61 |
| <i>Shorea robusta</i> | 42.2 | 26.8 | 2137.77 | 1004.75 | <i>Shorea robusta</i> | 67.1 | 25 | 5041.80 | 2369.65 |
| <i>Careya arborea</i> | 31.6 | 15.8 | 706.69 | 332.15 | <i>Shorea robusta</i> | 60 | 30 | 4837.54 | 2273.64 |
| <i>Shorea robusta</i> | 45.2 | 24.5 | 2242.04 | 1053.76 | <i>Pterospermum lanceaefolium</i> | 10.6 | 9 | 45.30 | 21.29 |
| <i>Shorea robusta</i> | 17.8 | 12.8 | 181.66 | 85.38 | <i>Anthocephalus chinensis</i> | 85 | 24 | 7766.93 | 3650.46 |
| <i>Shorea robusta</i> | 39.5 | 24 | 1677.28 | 788.32 | <i>Shorea robusta</i> | 41 | 19 | 1430.61 | 672.39 |
| <i>Semicarpous anacardium</i> | 33.5 | 13.4 | 673.59 | 316.59 | <i>Shorea robusta</i> | 12 | 9 | 58.05 | 27.28 |
| <i>Pterospermum lanceaefolium</i> | 25.5 | 15.4 | 448.54 | 210.81 | <i>Anthocephalus chinensis</i> | 80 | 26 | 7453.39 | 3503.09 |